

1 **A multi-instrumental approach for calibrating ~~two~~ real-time  
2 mass spectrometers using high performance liquid  
3 chromatography and positive matrix factorization**

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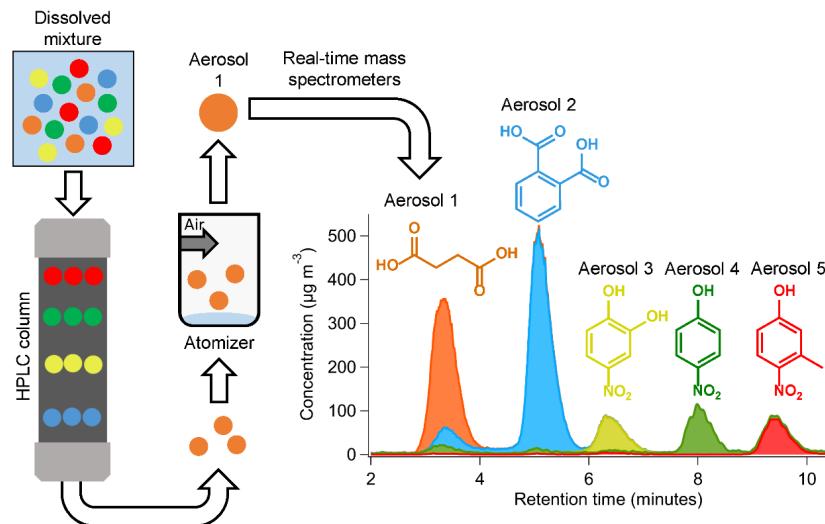
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10

11 **Abstract.** Obtaining quantitative information for molecular species present in aerosols from real-time mass  
12 spectrometers such as an extractive electrospray time-of-flight mass spectrometer (EESI) and an aerosol mass  
13 spectrometer (AMS) can be challenging. Typically, molecular species are calibrated directly through the use of pure  
14 standards. However, in some cases (e.g. secondary organic aerosol [SOA] formed from volatile organic compounds  
15 [VOCs]) direct calibrations are impossible, as many SOA species can either not be purchased as pure standards or  
16 have ambiguous molecular identities. In some cases, bulk from real-time soft-ionization aerosol instruments such as  
17 an Extractive Electrospray time-of-flight Mass Spectrometer (EESI) can be challenging, due to many individual  
18 species having different, and often hard to predict, sensitivities. Directly calibrating is time consuming and relevant  
19 standards are often hard to obtain. In addition, the molecular identities of many of the sampled species may be  
20 ambiguous. Bulk OA sensitivities are sometimes used to estimate molecular sensitivities. This approach is not  
21 sufficient for EESI, which measures molecular components of OA, because different species, but different types of  
22 OA, can have bulk sensitivities that vary by a factor of ~10–more than 30. Here, we introduce a method to obtain  
23 EESI calibration factors when standards are not available, and we provide a thorough analysis of the feasibility,  
24 performance, and limitations of this new technique. In this method, complex aerosol mixtures were separated with  
25 high performance liquid chromatography (HPLC) followed by aerosol formation via atomization. The separated  
26 aerosols were then measured by an EESI and an AMS, which allowed us to obtain sensitivities for some species  
27 present in standard and SOA mixtures. Pure compounds were used to test the method and characterize its  
28 uncertainties, and obtained sensitivities were consistent within  $\pm 20\%$  when comparing direct calibrations vs HPLC  
29 calibrations for a pure standard, and within a factor of two for a standard mixture. In some cases, species were not  
30 completely resolved by chromatography, and positive matrix factorization (PMF) of AMS data enabled further  
31 separation. This method should be applicable to other real-time MS techniques. Improvements in chromatography  
32 are possible that would allow better separation in complex mixtures. A system to separate the compounds present in

33 complex samples can enable their direct calibration. Here, high performance liquid chromatography (HPLC)  
 34 followed by aerosol formation via atomization was combined with online, 1 Hz measurements to calibrate the EESI  
 35 and a High Resolution Aerosol Mass Spectrometer (AMS) for compounds present in a secondary organic aerosol  
 36 (SOA) mixture. Pure compounds were used to test the method and characterize its uncertainties. Pure compound  
 37 calibration factors were consistent within  $\pm 20\%$  for direct atomization vs. HPLC separation, which is far superior to  
 38 the orders of magnitude sensitivity differences that are possible with EESI. For species that were not well separated  
 39 by chromatography, Positive Matrix Factorization (PMF) based on AMS spectra was used to test its ability to  
 40 separate overlapping species. In two test cases, further separation was achieved using PMF, but derived sensitivities  
 41 from direct and HPLC calibrations varied by up to a factor of 2.

42 TOC figure



43

44 **1 Introduction**

45 Atmospheric aerosols are a complex, and often poorly understood, component of Earth's atmosphere. Aerosols have  
46 significant effects on both human and ecosystem health, and are significant contributors to anthropogenic climate  
47 forcing (Dockery et al., 1996; Lighty et al., 2000; Lohmann et al., 2004; IPCC, 2013). Organic aerosol (OA) is a  
48 substantial component of global aerosol levels (Kanakidou et al., 2005; Zhang et al., 2007; Jimenez et al., 2009).  
49 Since the early 2000s an important instrument for measuring OA concentrations in real-time has been the Aerosol  
50 aerosol Mass-mass Spectrometer-spectrometer (AMS) (Jayne et al., 2000; Canagaratna et al., 2007) and its high-  
51 resolution version (HR-AMS) (DeCarlo et al., 2006). Soft-ionization aerosol mass spectrometers, such as the  
52 Extractive-extractive Electrospray electrospray Time-time-of-Flight flight Mass-mass Spectrometer-spectrometer  
53 (EESI-ToF-MS, EESI hereinafter), have more recently become important tools for obtaining more detailed OA  
54 speciation (Lopez-Hilfiker et al., 2014, 2019; Eichler et al., 2015).

55 EESI can detect individual molecular ions (referred to henceforth as either molecular ions or individual  
56 species, even if they may comprise several isomers) from the particle-phase with 1 s time resolution (Lopez-  
57 Hilfiker et al., 2019; Pagonis et al., 2021). EESI has been used to measure aerosols in urban areas (Qi et al., 2019,  
58 2020; Stefenelli et al., 2019; Kumar et al., 2022), in biomass burning (Qi et al., 2019; Pagonis et al., 2021), in  
59 cooking emissions (Qi et al., 2019; Brown et al., 2021), and for chamber studies of secondary OA (SOA) formation  
60 (Liu et al., 2019; Pospisilova et al., 2020). Many studies have illustrated the low detection limits, limited  
61 fragmentation, and other capabilities of the EESI; e.g. Lopez-Hilfiker et al. (2019) and Pagonis et al. (2021).

62 However, obtaining quantitative information for individual species from EESI measurements of complex  
63 mixtures of unknown species can be challenging. This is due to each species having different, and often hard to  
64 predict, sensitivities (Law et al., 2010; Lopez-Hilfiker et al., 2019; Brown et al., 2021; Wang et al., 2021). In  
65 addition, EESI measures molecular ions, but can in some cases cause fragmentation, such as due to loss of  $\text{HNO}_3$   
66 from nitrates (Liu et al., 2019). For an SOA mixture from a single precursor, the bulk sensitivity compared to SOA  
67 formed from a different precursor has been shown to vary by a factor of 15 or more (Lopez-Hilfiker et al., 2019).  
68 Different studies also show that the bulk sensitivity for OA formed from different emission sources, (e.g. cooking,  
69 biomass burning,) can vary by a factor of  $\sim 10$  (Qi et al., 2019; Stefenelli et al., 2019; Brown et al., 2021). For pure  
70 organic standards, the sensitivity can vary by a factor of 30 or more (Lopez-Hilfiker et al., 2019). Instead of directly  
71 measuring compound sensitivity, some groups use machine-learning (Liigand et al., 2020) or thermodynamic  
72 modeling (Kruve et al., 2014) to approximate instrument response factors for individual species. Other studies use  
73 bulk calibration factors for complex mixtures as an approximation for quantification (Tong et al., 2022).

74 Sensitivities can vary due to differences in analyte solubility (Law et al., 2010), EESI working fluid  
75 composition, sample composition, and different instrument conditions and settings, including polarity and changes  
76 in inlet pressure (Lopez-Hilfiker et al., 2019; Pagonis et al., 2021). Calibrating the EESI for individual species can  
77 be a challenging task, especially when standards are unavailable for most atmospheric oxidation products. In  
78 addition, OA from chamber experiments or field studies often contains unidentified molecular ions, or those whose  
79 species identity is ambiguous.

80 Several calibration methods have been applied to EESI. For example, direct calibrations were performed  
81 for many organic standards in Lopez-Hilfiker et. al. (2019), for 4-nitrocatechol (EESI $^-$ ) and levoglucosan (EESI $^+$ )  
82 in Pagonis et al. (2021) to track sensitivity during each aircraft flight, and levoglucosan for regular sensitivity  
83 tracking during an indoor cooking study (and several other compounds less frequently and bracketing the campaign)  
84 in Brown et. al. (2021). During research field studies, often only one or two species are calibrated frequently, and  
85 the rest are quantified using relative response factors measured less frequently (Qi et al., 2019; Brown et al., 2021;  
86 Pagonis et al., 2021).

87 A recent paper-study combined measurements from the Vocus Proton-Transfer Mass Spectrometer  
88 (Vocus), AMS, and EESI to measure speciated response factors without the need for standards. In that study, SOA  
89 was generated using an Oxidation-oxidation Flow-flow Reactor-reactor (OFR). Following SOA formation, the Vocus  
90 measured the gas phase species, and the AMS and EESI measured the bulk and speciated particulate phase,  
91 respectively. EESI response factors were obtained through comparison to decreasing gas-phase mixing ratios  
92 measured by the Vocus as they condensed to the particle-phase (Wang et al., 2021).

93 Another method for obtaining calibration information is Positive-positive Matrix-matrix Factorization  
94 factorization (PMF). PMF is a type of factor analysis that allows approximately apportioning of aerosol mass  
95 measured with online mass spectrometers and other instruments to atmospheric sources or level of oxidation (Zhang  
96 et al., 2005; Lanz et al., 2007; Ulbrich et al., 2009). To our knowledge, PMF has not been used with AMS data alone  
97 to obtain mass spectra and time series for individual molecular components. Separation with PMF alone would  
98 could be difficult for ambient or chamber experiment data, in part, since most compounds likely co-vary in time and  
99 thus would not be statistically resolvable (Craven et al., 2012). Direct calibrations have been conducted to generate  
100 high-resolution AMS mass spectra for individual species (Ulbrich et al., 2019). A combination of AMS and PMF  
101 has been used to obtain quantitative information for EESI bulk measurements or PMF factors (Qi et al., 2019, 2020;  
102 Kumar et al., 2022). PMF has also been used on a combined data set consisting of both EESI and AMS data (Tong  
103 et al., 2022).

104 To our knowledge, PMF has not been applied previously to AMS and EESI chromatographically-separated  
105 data. Running PMF on chromatographic data may be able to generate species-specific mass spectra and time series  
106 for compounds that cannot be obtained directly as pure standards. PMF has been applied in the past to gas  
107 chromatography mass spectrometry (GC-MS) data (Zhang et al., 2014, 2016; Gao et al., 2018), but not to High-high  
108 Performance-performance Liquid-liquid Chromatography-chromatography (HPLC) data, which is better suited for  
109 oxidized SOA species than GC, to our knowledge. AMS detection following HPLC separation has been conducted  
110 previously (Farmer et al., 2010) to explore AMS spectra of the separate compounds, but not for quantification.  
111 HPLC has not been previously combined with EESI or PMF, to our knowledge. Further, HPLC must be used here  
112 because the mass spectrometric detection needs to be much faster than the chromatographic time scale (on the order  
113 of seconds). Otherwise, this method is not applicable, and the different species separated by the chromatography  
114 would not be sufficiently resolved for speciated detection with the EESI and AMS.

115 Here, for the first time, we demonstrate a method combining High Performance Liquid Chromatography  
116 (HPLC), atomization, and detection by EESI, AMS, and sScanning Mobility-mobility pParticle sSizer (SMPS). The

117 method was validated by ~~running pure standards, standard mixtures, and separating a mixture of standards, and then~~  
118 ~~applied to~~ chamber SOA. The analyte peak~~s~~ measured with each instrument ~~were~~as integrated, and calibration  
119 factors for separated species were calculated for the EESI( $CF_x^E$ ). The AMS response factor ( $CF_x^A$ , or RIE-\*CE, the  
120 product of the relative ionization efficiency and collection efficiency) and the atomic oxygen to carbon (O:C) ratio  
121 for different analytes were quantified. EESI calibration factors ( $CF_x^E$ ) for individual compounds were determined  
122 and compared to literature values. In cases where ~~full peak separation via HPLC alone was not achieved did not fully~~  
123 ~~resolve all analytes~~, PMF was run on the ~~EESI and~~AMS mass spectral matrices to obtain further compound  
124 separation.

125 **2 Methods**126 **2.1 Chamber experiments and filter mass collection**

127 SOA was generated using the procedure of DeVault et. al. (2022). Briefly, chamber experiments were conducted in  
128 a ~~6.9 ( $\pm 0.5$ )~~ 8.0 m<sup>3</sup> Teflon chamber ([Claflin and Ziemann, 2018](#); [Bakker-Arkema and Ziemann, 2021](#))[Bakker](#)  
129 [Arkema and Ziemann, 2021](#)). The temperature (23 °C) and atmospheric pressure (0.83 atm) were constant.  
130 Ammonium sulfate seed was added to ~~a~~the humidified chamber (RH = ~~55~~ %), followed by β-pinene, which was  
131 evaporated from a heated glass bulb. In the dark, N<sub>2</sub>O<sub>5</sub> was added as the NO<sub>3</sub> source, from the sublimation of  
132 cryogenically-trapped solid N<sub>2</sub>O<sub>5</sub>. [During these experiments, ~ 372 - 1378 μg m<sup>-3</sup> SOA was made within the large](#)  
133 [reaction chamber. This material was collected on a filter for ~ 120 min at a flow rate of 14 L min<sup>-1</sup>. Following](#)  
134 [dissolution in solvent, ~ 16 - 56 μg of SOA was injected into the HPLC. Further discussion is included in Sect. S4.](#)

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135 The experiment was modeled after Claflin et. al. (2018).

136 Following SOA formation, a 0.45 μm Millipore Fluoropore PTFE filter was used to collect SOA. The  
137 [combined filter and aerosol was weighed](#) after aerosol collection. The [combined filter and aerosol](#) was  
138 exposed to minimal ambient air, and was always handled with artificial lighting turned off and outdoor blinds  
139 drawn. After weighing, each filter was extracted in 2 mL of HPLC grade ethyl acetate (EtAc) twice. The 4 mL  
140 aerosol extract/EtAc mixture was dried using pure N<sub>2</sub>. Once the EtAc was evaporated, the leftover material was  
141 dissolved in HPLC grade acetonitrile (ACN) and stored in a freezer at -23 °C (DeVault et al., 2022). The extract  
142 used here was the same as DeVault et. al. (2022), and was ~~one~~1 year old at the time of analysis. [That study showed](#)  
143 [that the DeVault et al. \(2022\) showed that this](#) SOA is composed entirely of acetal dimers, which are exceptionally  
144 stable. [Therefore, so](#) the SOA is unlikely to have changed [significantly](#) over this period.

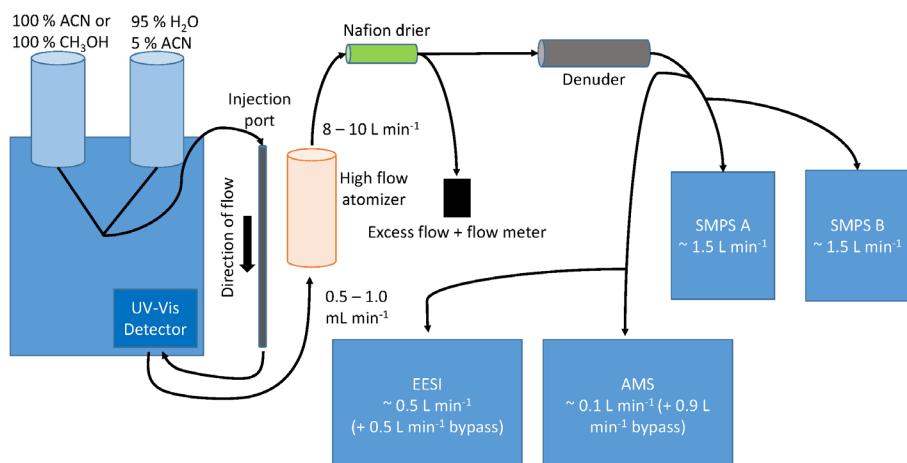
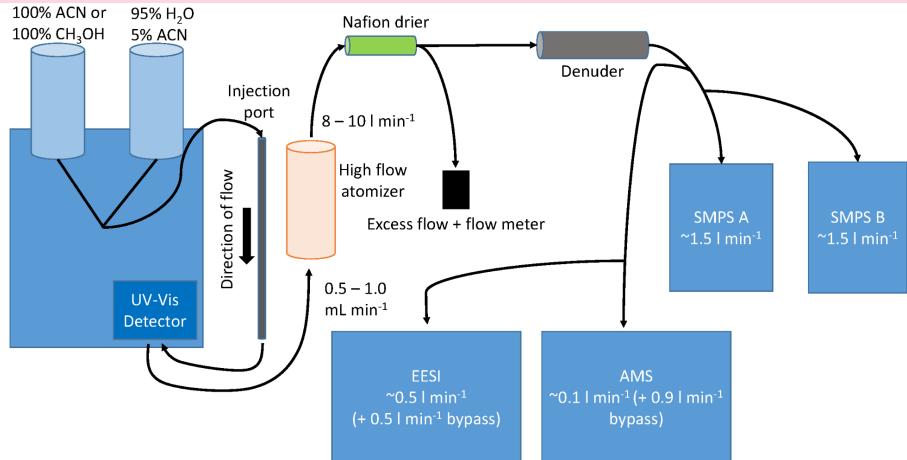
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145 **2.2 High Performance Liquid Chromatography (HPLC)**

146 HPLC separation was performed using a Shimadzu Prominence HPLC, coupled to a Zorbax Eclipse XDB-C18  
147 column (250 × 4.6 mm with 5 μm particle size). A Nexera X2 SPDM30A UV/vis photodiode array detector was  
148 used to generate absorbance chromatograms. The column stationary phase was designed for reverse mode, where  
149 smaller, more polar species had shorter elution times. Separated species were measured first at  $\lambda = 210$  nm and  $\lambda =$   
150 254 nm using an UV-Vis diode array detector with a reference wavelength of 300 nm. Separated chemical  
151 components then flowed into a ~~high-~~ flow Collison atomizer, forming droplets and then aerosols consisting solely of  
152 the SOA compounds after evaporating the HPLC solvent in a Nafion drier. The aerosols were then measured by a  
153 suite of instruments, shown in Fig. 1, and pictured in Fig. S1. Tubing delay times are also included in Table S1.  
154

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157 **Figure 1. HPLC schematic. Left, HPLC containing a column and a UV-Vis detector. Following separation, the column**  
158 **effluent was sent to an atomizer, dried, and the aerosol was detected by each of the instruments shown.**

159  
160 A maximum volume of 50  $\mu$ L ACN/aerosol mixture was injected into the column at once. At the beginning of each  
161 day, the HPLC solvent lines (HPLC grade acetonitrile and HPLC grade water) were flushed to remove any air  
162 bubbles that may affect elution. Following this, a clean cycle was run by injecting 50  $\mu$ L HPLC grade ACN into the  
163 reverse-phase column. This ensured previous HPLC run species did not contaminate new runs. The first run of the

164 day, post cleaning cycle, was a 4-nitrocatechol/4-nitrophenol mixture (dissolved in ACN). These species were well  
165 characterized by the particle-phase instruments and have measurable absorbances at the recorded UV wavelengths.

166 For each experiment, the mobile phase consisted either of an ACN/water mixture or an ACN/CH<sub>3</sub>OH/water mixture. The mixture varied in relative concentrations of each solvent over the course of each HPLC run.  
167  
168 Most experiments were started at 95% water/5% ACN (solvent mixture A). The mobile phase became less polar  
169 over time. For some systems, solvent B (pure acetonitrile) replaced solvent system A as time went on. For other  
170 systems, solvent C (pure methanol) was used. Each standard and/or SOA system was run under different  
171 conditions, depending on the separability of different components.

172 For the standard solution run, a mixture of solvent A and solvent B was used. Using a flow of 1.0 mL min<sup>-1</sup>,  
173 solvent B was increased from 0% to 35% in 1 minute, then 35%–40% for 5 minutes, followed by 40%–50%  
174 for 3 minutes, and 50%–100% for 2 minutes, this is also shown in Fig. S2a. For the β-pinene SOA extract, the  
175 flow rate was set to 0.5 mL min<sup>-1</sup>, and a mobile phase gradient started at 20% solvent C for 2 minutes, then  
176 increased at a rate of 6% min<sup>-1</sup> up to solvent C of 50%, followed by an increase of 3% min<sup>-1</sup> to a concentration of  
177 80% solvent C, then 0.75% min<sup>-1</sup> until 95% solvent C, held at 95% C for 20 minutes and increased by 1.7% min<sup>-1</sup>  
178 to 100%, following 10 minutes at 100% solvent B, shown in Fig. S2b (DeVault et al., 2022).

### 179 2.3 Standards for HPLC measurements

180 Two standard solutions of atmospherically relevant species were made for this study. Standard solution 1 contained  
181 0.4% (by mass) 3-methyl-4-nitrophenol, 0.2% phthalic acid, 0.5% 4-nitrophenol, 0.6% succinic acid, and 0.1% 4-  
182 nitrocatechol, dissolved in HPLC grade acetonitrile. Solution 2 contained 8 species: 0.3% phthalic acid (by mass),  
183 0.3% L-malic acid, 0.1% succinic acid, 0.3% citric acid, 0.3% levoglucosan, and 0.2% 4-nitrocatechol in HPLC  
184 grade acetonitrile. Source information and calculated saturation mass concentrations for all species are shown in  
185 Table S2.

186 Each species was chosen for its relevance in biomass, urban, or manufacturing processes. 3-methyl-4-  
187 nitrophenol, 4-nitrophenol, 4-nitrocatechol and levoglucosan are cyclic C<sub>6</sub> carbon species found in biomass burning.  
188 Succinic acid, L-malic acid, and phthalic acid are ~~non-eyelie~~ acids of secondary origin found in urban atmospheres.  
189 Citric acid is found in food and/or medicine. A critical property of these compounds is that they absorb in the UV-  
190 Vis, whereas most SOA does not. Nitrates and aromatics have strong absorbance and carboxylic acids have a very  
191 weak absorbance.

### 192 2.4 Aerosol Generation and Sampling System

193 The HPLC was coupled to particle phase measurements by using a high-flow Collison atomizer. First, a ~~t~~eflon line  
194 was attached to the waste port of the HPLC. The flow from the HPLC was 0.5–1 mL min<sup>-1</sup>, all of which was sent to  
195 the atomizer. The atomizer operated by first introducing pressurized compressed air (~20 psi) into a small chamber  
196 (473 mL jar). Perpendicular, sample flow at a rate of 0.5 or 1 mL min<sup>-1</sup> intersected the pressurized air. This led to  
197 the generation of particles of a consistent size distribution, and provided a total flow ranging from 8 to 10 L min<sup>-1</sup>.  
198 Instrument specific flows were measured daily.

199 Following atomization,  $\sim 10.4 \text{ L min}^{-1}$  of aerosol/solvent flow was sent through a Nafion dryer before  
200 being sent through an activated carbon denuder. This denuder is in a stainless steel,  $\sim 1$  inch diameter and 8 inch  
201 length tube, composed of activated carbon honeycomb cross-sections. Flow was then sent into each particle  
202 instrument. Solvent was efficiently removed ( $>99.0\%$ , Pagonis et. al. (2021)) using the carbon denuder.  
203 Acetonitrile (a solvent used in the HPLC system) was monitored using the EESI. If acetonitrile started to increase,  
204 the EESI denuder was regenerated. Denuder regeneration was typically only necessary after the first 4 h of each  
205 experiment.

206 Residence times in different parts of the system were estimated to enable synchronizing the aerosol  
207 instrument observations with and the measured UV-Vis absorbances. Calculations shown in Table S1 suggest that a  
208 delay of at least 41-40 seconds should be observed between the UV-Vis measurement and detection with the aerosol  
209 instruments, which is consistent with the measured delay. Retention times for EESI, AMS, and SMPS may differ  
210 from each other by 1-2 seconds, depending on the residence times in the tubing. In addition, bypass flows (shown  
211 in Fig. 1) were added to the EESI and AMS to reduce residence times in the tubing and thus particle losses or  
212 evaporation. These delay differences were handled by shifting instrument data by the delay times.

## 213 2.5 Description of particle measurements

### 214 2.5.1 Extractive Electrospray Time-of-Flight Mass Spectrometry (EESI)

215 The EESI uses a soft ionization technique that detects particle-phase analytes based on their solubility and proton  
216 affinity/adduct formation stability (Lopez-Hilfiker et al., 2019). Briefly, particle/gas sample flow was sent into  
217 the EESI source at  $\sim 0.5-1 \text{ L min}^{-1}$ , where gases are removed using a charcoal denuder ( $>99\%$  removal efficiency  
218 for acetic acid, when regenerated daily) (Tennison, 1998; Pagonis et al., 2021). The aerosol inlet for the instrument  
219 used in this study was pressure controlled (Pagonis et al., 2021), and was run at 766-575 mbar. While designed for  
220 aircraft applications, the pressure-controlled inlet provides better spray and signal stability as it shields the spray  
221 from small pressure perturbations from changes in upstream inlet flow conditions. This includes perturbations  
222 caused by such as switching between different sampling modes and plumbing pathways. The Here, the working  
223 fluid consisted of a mixture of 25% milli-Q water and 75% (by volume) HPLC grade methanol. The EESI was run  
224 in two polarity modes. The positive polarity mode (henceforth “EESI+”) contained 200 ppm of sodium iodide (NaI)  
225 (Pagonis et al., 2021). This working fluid generally forms Analyte-Na<sup>+</sup> adducts. The negative polarity mode (EESI-)  
226 was doped with 0.1% (by volume) formic acid (Chen et al., 2006; Gallimore and Kalberer, 2013; Pagonis et al.,  
227 2021). Species with a lower proton affinity than formic acid donate a proton and become negatively charged. This  
228 ionization mode is generally sensitive to acidic species that can readily donate a proton and become anionic.

229 For both polarities, a fused silica capillary (TSP Standard FS tubing, 50  $\mu\text{m}$  ID, 363  $\mu\text{m}$  OD) was used to  
230 transport working fluid solution from a pressurized (250-300 mbar above ambient) fluid bottle. Typical resolution  
231 at  $m/z$  150 was 4000, and mass spectra were saved every second.

232 The mass concentration of a species ( $\mu\text{g m}^{-3}$ ) can be quantified from its EESI signal ( $I_i$ , ion counts  $\text{s}^{-1}$ ) as  
233 (Lopez-Hilfiker et al., 2019):

234  $Mass_x = I_x \left( \frac{MW_x}{RF_x} \right) \cdot \frac{1}{F}$  (1)

235  $MW_x$  is the molecular weight of species  $x$ ,  $F$  is the flow rate (in  $L \text{ min}^{-1}$ ), and  $RF_x$  is the combined response factor.  
 236 There are representing fundamental parameters for EESI signal which can be found are described further in Lopez-  
 237 Hilfiker et. al. (2019). Here, we define a new variable, EESI calibration factor ( $CF_x^E$ , in  $\mu\text{g m}^{-3} \text{ counts}^{-1} \text{ s}$ ), such that

238  $Mass_x = I_x \cdot CF_x^E$  (2)

239 Generally,  $CF_x^E$  is directly determined by direct calibrations with standards, when possible. Here,  $CF_x^E$  was  
 240 determined by either direct calibrations using either commercially available standards or HPLC-separated analytes.  
 241 Calibration factors are reported as absolute values (in units of  $\text{counts s}^{-1} \mu\text{g}^{-1} \text{ m}^{-3}$ ) and also relative to 4-nitrocatechol  
 242 for EESI- and levoglucosan for EESI+ (unitless).

243 **2.5.2 High Resolution Aerosol Mass Spectrometer (HR-AMS)**

244 A high-resolution time-of-flight aerosol mass spectrometer (hereinafter AMS) (DeCarlo et al., 2006; Canagaratna  
 245 et al., 2007) was used to obtain 1 Hz chemical composition for organic aerosol (OA) and nitrate aerosol ( $\text{pNO}_3$ ). The  
 246 AMS was run with an inlet flow of  $0.1 \text{ L min}^{-1}$ , and a bypass flow of  $\sim 1.3-4 \text{ L min}^{-1}$ . The AMS was run  
 247 exclusively in “fast mode” (Kimmel et al., 2011; Nault et al., 2018), and size distributions were not recorded. AMS  
 248 backgrounds were measured for 6 seconds every 52 seconds. Outside of HPLC runs, the AMS background was <  
 249  $0.1 \mu\text{g m}^{-3}$ . Between eluting peaks Additional backgrounds were taken, in part to test for solvent residue and / or  
 250 residual influence from the previous HPLC runs, were taken during the times where no peaks were eluting, and  
 251 These backgrounds were generally remained <  $2 \mu\text{g m}^{-3}$  for both the AMS and the SMPSs. The detection limit (DL)  
 252 and limit of quantification between eluting peaks was  $0.7 \mu\text{g m}^{-3}$  and  $2.2 \mu\text{g m}^{-3}$ , respectively, suggesting that  
 253 background subtracted concentrations above  $2.2 \mu\text{g m}^{-3}$  can be accurately measured. The latter were conducted by  
 254 flowing the sampler air through a particle filter. AMS data was analyzed in the ToF-AMS analysis software (PIKA  
 255 version = 1.25F, Squirrel = 1.65F) (DeCarlo et al., 2006; Sueper, 2023) within Igor Pro 8 (Wavemetrics, Lake  
 256 Oswego, OR). The When AMS sensitivities were not obtained from direct measurements, the AMS OA relative  
 257 ionization efficiency (RIE) and collection efficiency (CE) were assumed to be 1.4 ( $\text{OA}_{\text{default}}$ , (Canagaratna et al.,  
 258 2007)) and 1, respectively. The AMS  $\text{NO}_3$  RIE \* CE ( $\text{NO}_3$ ,  $\text{default}$ ) was assumed to be 1.1 (Canagaratna et al., 2007).  
 259 Data herein is reported in  $\mu\text{g m}^{-3}$ , using Boulder pressure ( $P = 830 \text{ mbar}$ ) and average lab temperatures ( $\sim 20^\circ\text{C}$ ).

260 Here, the quantification of different particle-phase species that have been separated by HPLC (and thus are  
 261 mostly in single component particles) is assessed for the AMS. This is a function of  $\text{RIE}_x \cdot \text{CE}_x$  (a.k.a. “AMS  
 262 response factor”, or  $CF_x^A$ ) for a species  $X$ . Direct AMS calibration has been reported for many OA species (Slowik  
 263 et al., 2004; Dzepina et al., 2007; Jimenez et al., 2016; Xu et al., 2018; Nault et al., 2023). An RIE of 1.4 is typically  
 264 applied to ambient organic aerosols (Canagaratna et al., 2007), which has been shown to perform well in most  
 265 outdoor intercomparisons (Jimenez et al., 2016; Guo et al., 2021). Laboratory measurements typically require

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266 specific calibrations, as RIE can be higher for some compounds and mixtures (Jimenez et al., 2016; Xu et al., 2018;  
267 Nault et al., 2023). CE can vary considerably, from  $CE = 0.15$  to a  $CE = 1$  (Docherty et al., 2013).

268 The material densities of the known standards were determined by running the AMS in PToF mode and  
269 calculating the density as  $d_{va}/d_m$ , where  $d_{va}$  is the aerodynamic vacuum diameter, and  $d_m$  is the SMPS measured  
270 mobility diameter (DeCarlo et al., 2004). Calculated densities are shown in table S2. For the unknown species  
271 present in the SOA, densities were estimated using the atomic ratio of oxygen plus nitrogen to carbon ( $[O+N]:C$ )  
272 and H:C, as demonstrated in Day et. al. (Day et al., 2022), which builds upon the method of Kuwata et. al. (Kuwata  
273 et al., 2012) which did not account for nitrate content. The O:C ratio attributed to the non-nitrate OA was calculated  
274 per Canagaratna et. al. (2015). The organic nitrate contribution was quantified per Day et. al. (2022). All nitrate here  
275 was assumed to be from organic nitrate functional groups, as the aerosol studied here likely contained little  
276 inorganic nitrate. For the density calculation, the total nitrate was multiplied by the ratio of the molecular weights of  
277  $NO_2:NO_3$  (46/62) and converted into a molar concentration using the molecular weight of  $NO_2$  (46 g mol<sup>-1</sup>). Only  
278 the  $NO_2$  functionality was included for the density calculation, since the nitrate oxygen bonded to the carbon is  
279 expected to typically be included as part of the standard AMS OA O:C estimation (Farmer et al., 2010). Carbon was  
280 also converted into a molar concentration using the molecular weight (12 g mol<sup>-1</sup>). That organic nitrogen to organic  
281 carbon ratio was added to the standard AMS OA O:C ratio to obtain the organic nitrate-corrected  $[O+N]:C$  ratio.

282 For isolated peaks that contained organic nitrate, the organic nitrate ( $NO_3$ ) concentration was added to the  
283 AMS OA to get the total measured AMS mass. The SMPS mass was then compared to the AMS mass calculated  
284 with the default  $CF_x^A$ , and the correct  $CF_x^A$  was determined with Eq. 3 (further details in Sect. 2.7).

$$285 \quad CF_x^A = \frac{OA_{default} + NO_3, default}{SMPS \text{ mass}} \quad (3)$$

286 For HPLC peaks composed of multiple species (like in the  $\beta$ -pinene SOA sample), the average  $CF_x^A$  was calculated  
287 by adding the average  $NO_3$  contribution (~5%) to the measured AMS OA contribution (Fig. S3). This  $CF_x^A$  was  
288 then applied to the AMS PMF organic chromatographic time series, in order to determine  $CF_x^E$ . For species not  
289 containing any nitrate, the  $NO_3, default$  was set to 0.

290 We note that some recent work has suggested that the sensitivity of organic nitrate functional groups may  
291 be lower than for ammonium nitrate (for which the nitrate is calibrated by default in AMS data processing). Thus, a  
292 correction of ~62/46 may be more appropriate here for computing nitrate functional group mass concentrations  
293 (Takeuchi et al., 2021). However, due to the small nitrate contribution overall, such a correction is-was not applied.

#### 294 2.5.3 Scanning Mobility Particle Sizer (SMPS)

295 Two SMPSs were run with a 20 second offset during HPLC experiments (consisting of all TSI, Inc components) in  
296 order to improve the time resolution of the total particle volume measurement. For both SMPSs, a 3081 differential  
297 mobility analyzer (DMA) was run with a 3080 Electrostatic-electrostatic classifier. Each was coupled with either a  
298 3776 condensation particle counter (CPC) (referred to as SMPS A) or a 3775 CPC (SMPS B). Both systems were  
299 run in the CPC “high flow” mode. Sample flow rates were nominally set to 1.5 4L min<sup>-1</sup>, but the actual (measured

300 flow) was 1.43 and 1.49  $\mu\text{L}$   $\text{min}^{-1}$  for the 3776 and 3775, respectively. DMA sheath flows were set to 6.0  $\mu\text{L}$   $\text{min}^{-1}$ .  
301 Data were compared to that acquired in a reference mode, with a sample flow of 0.3  $\mu\text{L}$   $\text{min}^{-1}$ , a sheath flow of 3.0  $\mu\text{L}$   $\text{min}^{-1}$ , and 120 s scans. Testing was done to ensure that number and volume distributions and integrated  
302 concentrations matched between the reference and fast scanning modes, shown in Fig. S4 and discussed in depth in  
303 Sect. S3. The SMPSs were also run concurrently during an HPLC run to confirm that data from both instruments  
304 matched (Fig. S5). Overall, the SMPSs in the reference and fast modes agreed within 10 %. Flows were measured  
305 every day, and delay times (from the SMPS inlet to the CPC detection, which affect sizing) were calculated when  
306 changes in plumbing were made. Further details on SMPS delays can be found in Table S3.

### 308 **2.5.4 Direct Calibration Procedure**

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309 Direct calibration refers to the standard method of generating monodisperse aerosol from a calibrant solution with a  
310 Collison atomizer (TSI model 3076) drying with a Nafion dryer, size selecting at 275 nm with a TSI 3080  
311 electrostatic classifier / 3081 DMA, removing double charged particles with an impactor, measuring the particle  
312 concentration with a 3775 CPC, and measuring with the EESI and / or AMS. The EESI and AMS sensitivities were  
313 obtained by comparing their signals to the particle mass calculated from the known particle volume, estimated  
314 density, and CPC particle concentration.

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### 315 **2.6 Positive Matrix Factorization (PMF)**

316 Positive Matrix Factorization (PMF) (Paatero and Tapper, 1994; Paatero, 1997) is a bilinear deconvolution model  
317 that relies on the assumption of mass balance with components with constant spectral profiles. Briefly, time series  
318 for signals at individual  $m/z$ 's are entered into a two-dimensional matrix with  $m$  rows (points in time) and  $n$  columns  
319 ( $m/z$ 's) (Ulbrich et al., 2009; Kumar et al., 2022). PMF works to minimize the squared weighted residuals between  
320 the measured and reconstructed matrices, producing multiple potential solutions that could explain different  
321 chemical or physical sources in a given data set, along with the total residual of each solution.

322 The model is solved using PMF2 (Paatero, 2007) and the multilinear engine, developed by Paatero et. al.  
323 (1999), run from the PMF Evaluation tool ("PET") software v3.08 in Igor Pro v8 (Wavemetrics, Lake Oswego, OR).  
324 Choosing the best PMF solution always has a subjective component, as it is usually impossible to know the  
325 "correct" number of factors that completely capture a complex data-set (Ulbrich et al., 2009). Several methods can  
326 be used to assess the validity of a given solution. First, the Q-value (Q), which is the total sum of the error-weighed  
327 square residuals for a data set, is used.  $Q_{\text{exp}}$  is the expected value of Q if all residuals are due to random errors with  
328 the estimated precision at each point. If the individual data points in a solution are fit so that the residuals are  
329 consistent with random noise, then  $Q/Q_{\text{exp}} \sim 1$ . Note that this also requires accurate estimation of the precision  
330 (random error) in the entire data matrix. In some situations, PMF cannot explain a data set within an acceptable  
331 error. In these situations,  $Q/Q_{\text{exp}} >> 1$ . All solutions here have  $Q/Q_{\text{exp}} \leq 1$ .

332 The second criteria for picking the best PMF solution is by exploring the time series and mass spectra for a  
333 given solution for different approximate rotations (FPEAK values) (Lee et al., 1999; Lanz et al., 2007; Ulbrich et al.,  
334 2009). Simply, PMF rotations are non-unique solutions that are represented across multiple factors. In a real-world

335 example, a source profile (for example, biomass burning OA), might split across multiple PMF factor's time series  
336 and/or mass spectra, despite only being from a singular source. Factor splitting can sometimes reduce residuals, and  
337 mathematically may appear as a more correct solution for a particular dataset. This is where the user must  
338 thoroughly assess different solutions, specifically those with  $Q/Q_{\text{exp}} \leq 1$ .

339 PMF solutions chosen here are based on the above criteria and a third: the time series of the residuals. In a  
340 chromatogram, the shape of the peaks ~~are~~ is generally known. Here, ~~4~~ four different instruments generate unique  
341 chromatograms: UV-Vis, ~~HR~~AMS, EESI, and the SMPSs. Thus, across those four instruments, the shape of the  
342 chromatogram was fairly well constrained. When choosing solutions here, the shape of the chromatogram was  
343 compared to the time series of the residuals. If the residuals showed significant peaks, then that was an indicator that  
344 not enough factors were used to represent the complete chromatogram and all of the factors therein.

345 The  $m \times n$  matrix for AMS data was generated for HR ions using the PMF export option in the PIKA data  
346 analysis software. Briefly, unit mass and high resolution AMS data were first fit as described in Sect. 2.5.2. After  
347 confirming that all ions of interest were well fit, the organic data was exported into an  $m \times n$  matrix (both signal  
348 and precision matrices). Any HR ions not associated with the following families: C<sub>x</sub>, CH, CHO<sub>l</sub>, and CHO<sub>gtl</sub> were  
349 removed, as NO<sub>3</sub> was not included in the PMF input, and the included families were the only measured ions with  
350 substantial signal during the experiments included here. PMF was run from 1 - 20 factors. Rotations (FPEAKS)  
351 were enabled, ranging from -1.0 to 1.0, in steps of 0.2.

### 352 2.7 Calculating calibration factors for species using the multi-instrumental method

353 For unknown species (or known species with an unknown AMS response factor) the following method was used to  
354 obtain EESI and AMS calibration factors:

- 355 1. Calculation of composition-dependent density using the measured elemental composition or  $d_{\text{va}}/d_m$   
356 measured densities [from AMS and SMPS data](#).
- 357 2. SMPS size distributions are fit with a lognormal curve, and integrated volume concentrations are obtained.
- 358 3. SMPS integrated volume time series were multiplied by the density, to produce the reference mass  
359 concentration time series.
- 360 4. The high-time-resolution AMS OA and NO<sub>3</sub> time series are obtained for an assumed RIE \* CE = 1.4  
361 (OA<sub>default</sub>) and RIE \* CE = 1.1 (NO<sub>3, default</sub>).
- 362 5. The SMPS mass concentration time series and the AMS OA+NO<sub>3</sub> time series, for an individual  
363 chromatographic peak, are fit with a Gaussian distribution
- 364 6. The AMS and SMPS Gaussian distributions are integrated ( $\mu\text{g m}^{-3} \text{ s}$ ).
- 365 7. The  $CF_x^A$  was obtained using the ratio of the integrated SMPS to the integrated AMS time series fits (Eq.  
366 3).
- 367 8. The time series for the EESI  $m/z$  was fit with a Gaussian and integrated along the retention time.
- 368 9. The integrated [Gaussian](#) for the EESI  $m/z$  was divided by the integrated AMS (OA+NO<sub>3</sub>, after AMS  
369 calibration by the SMPS) or SMPS [gaussians](#) [Gaussians](#) to obtain  $CF_x^E$  (counts  $\text{s}^{-1} \text{ m}^{-3} \mu\text{g}^{-1}$ ).

371 In step 9, the SMPS was used as the EESI reference for calculating  $CF_x^E$  when the analytes were resolved from  
372 chromatography alone. As discussed for the mixtures shown in Sect. 3.1, 3.2, and 3.3, we never obtained complete  
373 chromatographic separation. In cases of overlapping analytes, the SMPS used here does not have the time resolution  
374 to be used as the EESI reference. Instead, we referenced the EESI to the AMS by first calibrating the total AMS  
375 signal to the total SMPS signal for mixed peaks. We then used PMF results for the corrected AMS data and  
376 compared individual AMS PMF factors time series to EESI time series to calculate  $CF_x^E$ .

377 **3 Results**378 **3.1 Mass Balance of the Analyte in the Experimental System**

379 There was substantial plumbing between the injected sample and the instruments measuring the analyte, where  
 380 losses can occur (Fig. 1, Table S1). ~~In order to~~ To better understand the experimental system, the mass flux was  
 381 calculated using the known, injected mass as well as the tubing diameters, lengths, and flow rates, as shown in Fig.  
 382 2.

383

384

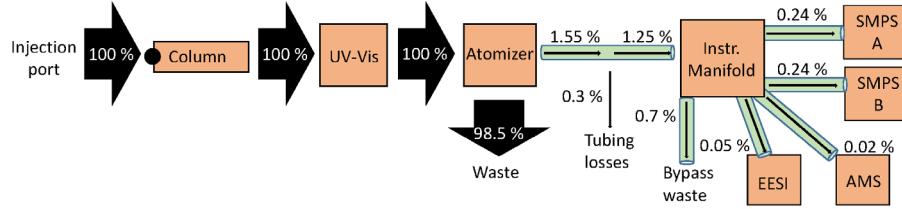
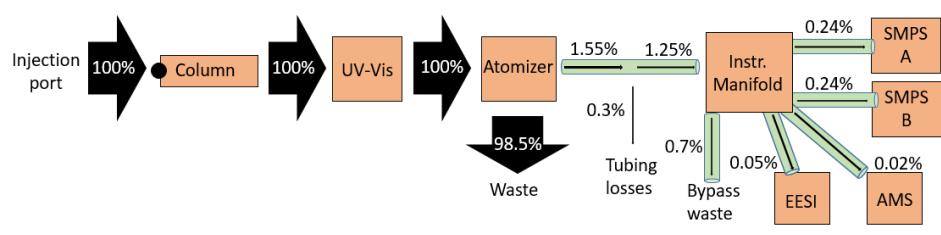
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386 **Figure 2. Mass flux across the multi-instrumental setup.** Arrows are sized by the percentage of analyte mass, which is  
 387 included alongside each arrow. EESI and AMS have bypass lines (represented as the total by 0.7% bypass waste).  
 388 Percentages shown are for the actual measured mass percent. Tubing details are also included in Fig. 1.

389

390 ~~By injecting a known amount of sample into the HPLC column, we were able to account for all the mass that was~~ allowed us to track all  
 391 the measured mass by the four instruments sampling. As shown in Fig. 2, all of the injected mass was analyzed by  
 392 the UV-Vis spectrometer, but only a small fraction of it was analyzed (0.55%) by the online instruments. There was  
 393 substantial fluid loss at the atomizer, which is thought to account for the bulk of the mass leaving the HPLC. The  
 394 EESI and AMS measure the least mass, due to their low flow rates ( $0.28 \text{ mL min}^{-1}$  and  $0.1 \text{ mL min}^{-1}$ , respectively).  
 395 Of the mass that exited the atomizer,  $\sim 20\%$  was lost in the tubing ( $\sim 10 \text{ m}$ ,  $1/4" \text{ I-D}$ ) to the aerosol sampling  
 396 manifold (represented as 0.3% of total in Fig. 2). Overall, the efficiency in sampling the injected mass with the  
 397 online instruments was very low with this system, primarily due to the atomization process. In SOA extracts that are  
 398 highly concentrated, this is not a major problem. However, application of this method to lower concentration  
 399 samples would benefit from use of a lower-flow liquid chromatography method and a more efficient atomizer.

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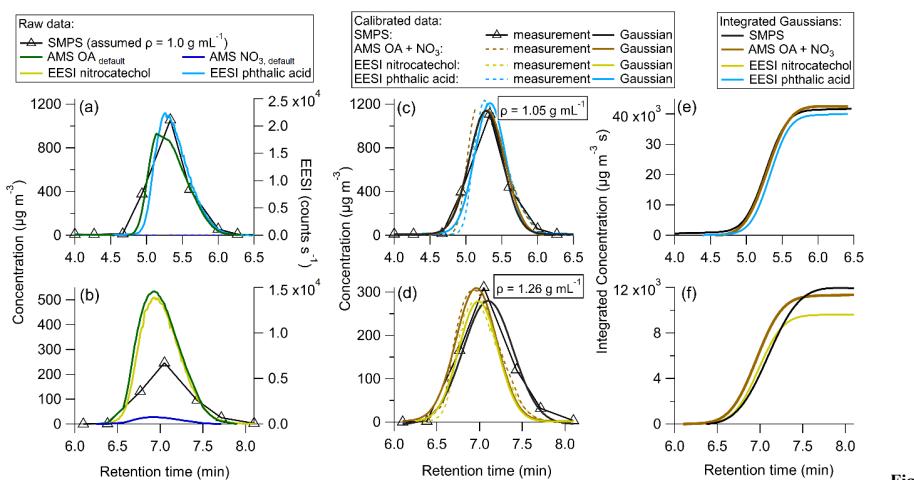
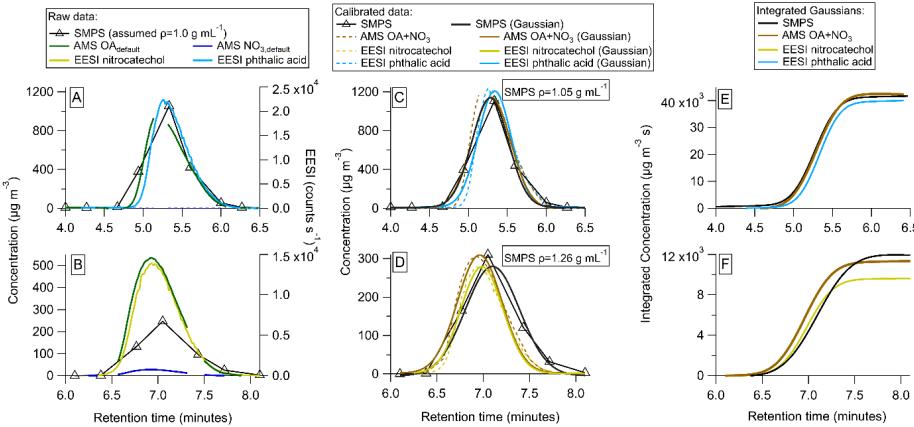
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401 **3.2 Application of multi-instrumental method and PMF for standard species' calibrations**

402 **3.2.1 Cross comparison between directly calibrated one-component chromatographic standards vs. multi-  
403 instrumental method**

404 In order to test the efficacy of the proposed method, two solutions were made containing one standard each (e. either  
405 phthalic acidand or 4-nitrocatechol). These species were first calibrated directly in order to obtain  $CF_x^E$  and  $CF_x^A$  as  
406 described in Sect. 2.5.4. Direct calibration hereout refers to the standard method of generating monodisperse aerosol  
407 from a calibrant solution with a Collision atomizer (TSI model 3076) drying with a Nafion dryer, size selecting at  
408 275 nm with a 3080 electrostatic classifier / 3081 DMA, removing double charged particles with an impactor,  
409 measuring the particle concentration with a 3775 CPC, and measuring with the EESI and/or AMS. Then, each  
410 solution was injected into the HPLC to generate isolated chromatograms (Fig. 3).

411



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Figure

414 3. Single standard calibrations for (Aa) uncalibrated data collected during a single standard (HPLC data for phthalic  
 415 acid) HPLC run, (Bb) raw data from a nitrocatechol HPLC run uncalibrated HPLC data for 4-nitrocatechol, (Cc)  
 416 calibrated phthalic acid data (using the monodisperse calibration factors) HPLC phthalic acid data calibrated using the  
 417 sensitivity derived from the direct calibration, (Dd) HPLC 4-nitrocatechol data calibrated using the sensitivity derived  
 418 from the direct calibration calibrated nitrocatechol data, (Ee) integrated Gaussian peaks from (Cc), and (Ff) integrated  
 419 Gaussian peaks from (Dd).

420

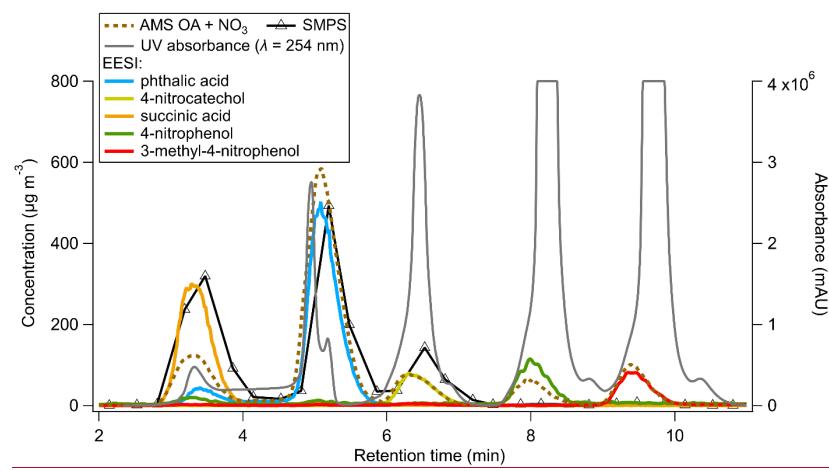
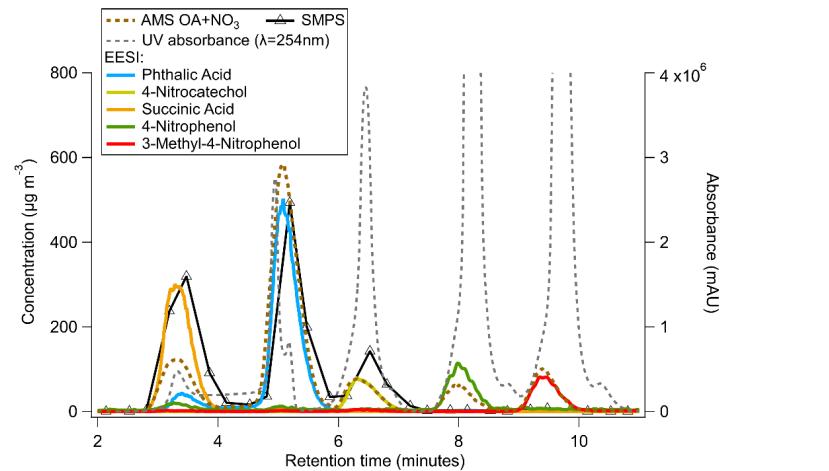
421 In Fig. 3a, the uncalibrated background-subtracted data is shown. Phthalic acid contains no nitrate moiety, so AMS  
 422 NO<sub>3</sub> was 0. Fig. 3b shows the raw data for 4-nitrocatechol. Due to the nitro group, AMS NO<sub>3</sub> is added to AMS OA  
 423 to obtain the total mass measured by the AMS. If the method was followed as described in Sect. 2.7, the raw data

424 would be fit with Gaussian curves and integrated, in order to produce  $CF_x^E$  and  $CF_x^A$  for each species. However, in  
425 this test study,  $CF_x^E$  and  $CF_x^A$  are already known [through direct calibrations discussed in Sect. 2.5.4](#).

426 Figure 3c shows the [HPLC phthalic acid peak with the direct calibration factor applied directly calibrated](#)  
427 [\(as opposed to the multi-instrumental approach calibrated\)](#) data for phthalic acid. It is clear that the AMS, EESI, and  
428 SMPS data line up well, indicating that the multi-instrumental approach produces very similar  $CF_x^E$  and  $CF_x^A$  as the  
429 direct calibrations. Fig. 3d echoes this, showing good overlap across each instrument for 4-nitrocatechol.

430 Figures 3e and 3f show the integrated, calibrated Gaussian curves. If the multi-instrumental method worked  
431 as well as direct calibrations, the maximum integrated values would be expected to be the same for each instrument.  
432 For phthalic acid, the instruments agree within 6 %, with the EESI showing the largest deviation from the other  
433 instruments. For 4-nitrocatechol, this difference is 20 %, and again the EESI is the farthest from the other  
434 instruments. Such discrepancies could be due to changes in EESI sensitivity, which may be driven by the different  
435 solvents used for calibration (water for direct calibrations, and a mixture of acetonitrile and water for the multi-  
436 instrumental method). It could also be due to the high concentrations of each solute, which may change  $CF_x^E$   
437 slightly.

438 Following method validation through comparison between direct calibrations and the multi-instrumental  
439 calibration method, a mixture containing five standards (phthalic acid, 4-nitrocatechol, succinic acid, 4-nitrophenol,  
440 and 3-methyl-4-nitrophenol) was run through the HPLC column (Fig. 4). Like above, each species was first  
441 calibrated directly, in order to compare the direct calibration values vs. the multi-instrumental calibration method for  
442 a more complex chemical system.



446 **Figure 4.** Time series of UV absorbance (milli-absorbance units) and AMS, EESI, and SMPS mass concentrations for a  
447 mixed-solution standard HPLC run.

449 In Fig. 4, succinic acid was the first peak to elute from the HPLC column, from ~2.5–4.0 minutes. The EESI and  
450 SMPS data match well, but the AMS data is lower by a factor of ~2. This is potentially driven by the phthalic acid/  
451 succinic acid co-elution (as evidenced by the EESI). The  $CF_x^A$  for both species is shown in Table 1.  $CF_x^A$  differ  
452 substantially, and an internal mixture of aerosols containing succinic acid and phthalic acid may result in a larger  
453 AMS bias (as  $CF_{Succinic Acid}^A$  and  $CF_{Phthalic Acid}^A$  differ significantly) than the EESI (where we measured molecular  
454 ions) or the SMPS (as the density of phthalic acid and succinic acid are similar, table S2).

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457 Table 1. Calibration factors for resolved (or mostly resolved) standard species.  $CF_x^E$  values are reported in counts  $s^{-1} \mu g^{-1}$   
 458  $m^3$  and the relative EESI calibrations factors ( $CF_x^E / CF_{nitro}^E$  (EESI-) or  $CF_x^E / CF_{levo}^E$  (EESI+)), and the AMS calibration  
 459 factors ( $CF_x^A$ ) are unitless values.

Species	Direct calibration $CF_x^E$ (counts $s^{-1} \mu g^{-1} m^3$ )	Multi-instr. calibration $CF_x^E$ (counts $s^{-1} \mu g^{-1} m^3$ )	Direct calibration $CF_x^E / CF_{nitro}^E$ (EESI-) or $CF_x^E / CF_{levo}^E$ (EESI+)	Multi-instr. calibration $CF_x^E / CF_{nitro}^E$ (EESI-) or $CF_x^E / CF_{levo}^E$ (EESI+)	Direct calibration $CF_x^A$ (unitless)	Multi-instr. $CF_x^A$ (unitless)
4-nitrocatechol (EESI-)	$44.1 \pm 5.0$	23	$1.0$	1	$1.962.0 \pm 0.17$	$1.051$
4-nitrocatechol (EESI+)	-	18	-	0.020	-	-
Succinic <u>Acid</u> (EESI-)	$30 \pm 4.0$	22	0.68	0.98	$1.6 \pm 0.10$	0.52
Succinic <u>Acid</u> (EESI+)	-	26	-	0.029	-	-
Phthalic <u>Acid</u> (EESI-)	$18.4 \pm 2.8$	18	0.41	0.82	$0.79 \pm 0.070$	1.0
Phthalic <u>Acid</u> (EESI+)	-	620	-	0.68	-	-
4-nitrophenol (EESI-*)	$1.6 \pm 0.57$	26	0.036	1.2	$0.59 \pm 0.050$	5.9
3-methyl-4-nitrophenol (EESI-*)	$5.8 \pm 4.0$	42	0.14	1.9	$0.90 \pm 0.10$	8.0
Levoglucosan (EESI+)	$200 \pm 10$	900	$1.0$	$1.0$	$0.45 \pm 0.06$	-

460 \* The reported values here are highly uncertain due to differences in evaporation for each instrument

461

462 Phthalic acid elutes as two isomers, with the largest eluting between 4 and 6 minutes. All three instruments match  
 463 well. 4-nitrocatechol was next, and showed very good agreement between the EESI and AMS, but a factor of  $\sim 2$   
 464 difference between the SMPS and AMS/EESI. The exact cause for this discrepancy is unknown.

465 4-nitrophenol and 3-methyl-4-nitrophenol both match well between the EESI / AMS, but the SMPS  
 466 concentration is a factor of 20 less than the other two instruments. The likely explanation is that 4-nitrophenol and 3-  
 467 methyl-4-nitrophenol are volatile (table S2). Compared to succinic acid,  $>90\%$  of these species evaporated from  
 468 injection to detection by the EESI / AMS. The SMPS measurement is slower than the other instruments, and dilutes

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469 the incoming aerosol by a factor of 4 inside the DMA column. The AMS and EESI measurements are faster and do  
470 not dilute the incoming aerosol. Due to these differences, nearly all of the injected mass evaporated in the SMPS.

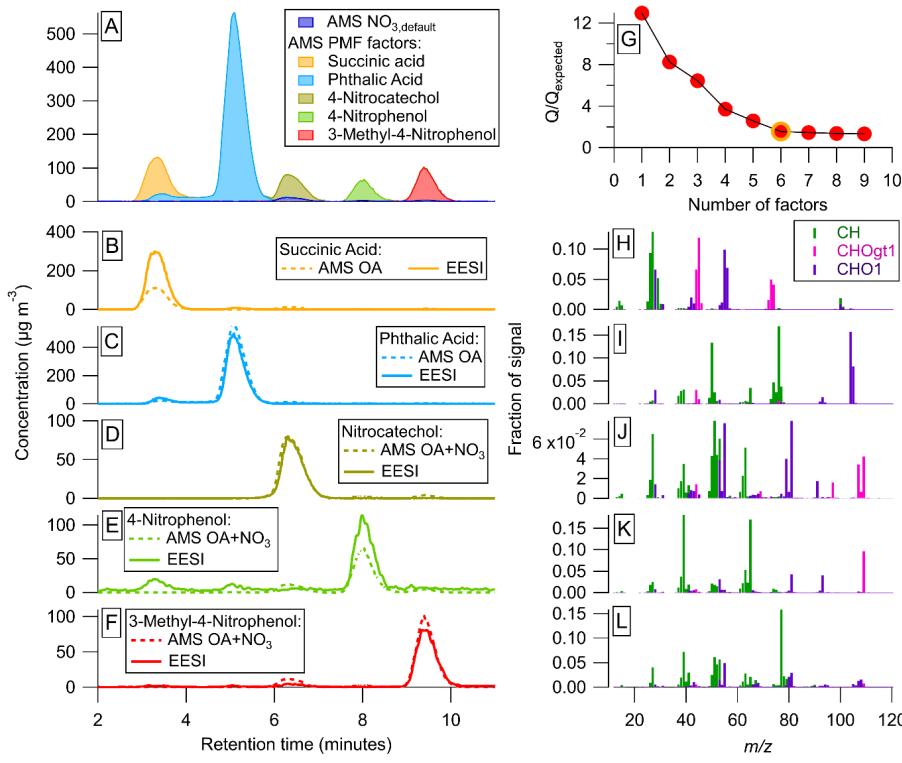
471 This suggests that volatile species (where  $C^* \gg OA$ ) are not able to be calibrated for by this method. Evaporation  
472 would also likely occur during direct calibrations, but to a lesser degree due to the higher pure-<sub>2</sub> species OA  
473 concentrations.

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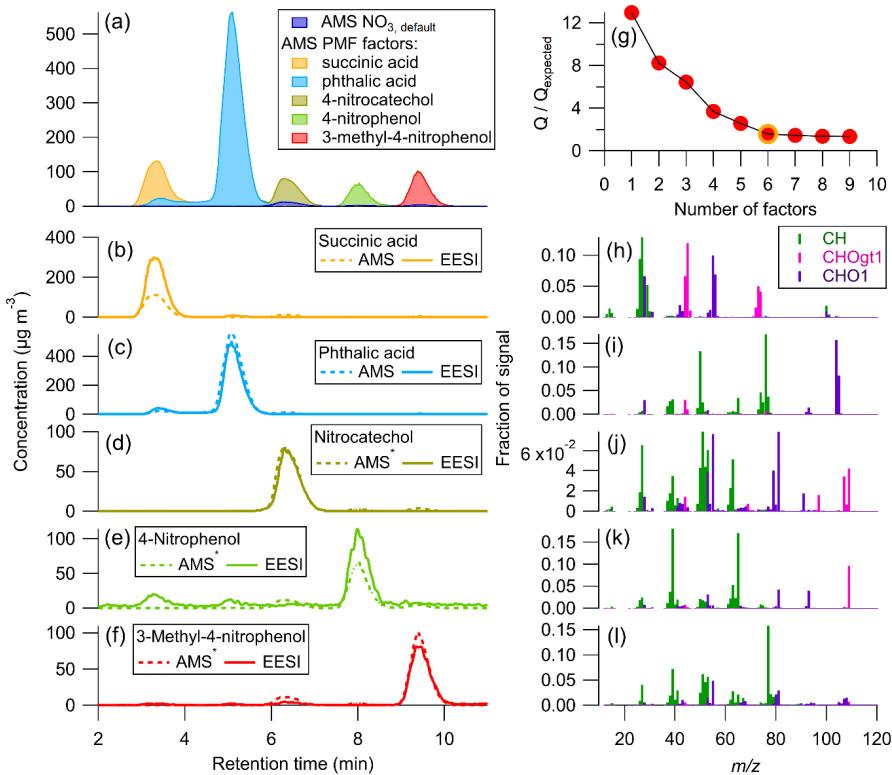
474 **3.2.2 Combined application of the multi-instrumental calibration method and PMF on two mixed standards  
475 solutions**

476 PMF was combined with the multi-instrument calibration method to better separate [the AMS data for](#) succinic acid  
477 and phthalic acid, which overlap in Fig. 4. The results of applying PMF to the AMS data is shown below in Fig. 5<sub>2,7</sub>  
478 [A 6-factor solution was chosen \(Fig. 5g\).](#)

479



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481  
482 **Figure 5.** Time series for the AMS PMF solution, (Aa) stacked plot of each factor and AMS NO<sub>3</sub>, (Bb)-(Ff) PMF factor  
483 with  $CF_x^A$  applied to individual species, along with EESI concentrations. (Gg)  $Q/Q_{\text{expected}}$ -vs. number of PMF factors,  
484 chosen solution circled in yellow. (Hh)-(Ll) mass spectra (colored by associated AMS HR family) for each AMS PMF  
485 factor. **A 6 factor solution was chosen, with only 5 factors plotted here. The remaining factor was attributed to the**  
486 **background signal, and was < 2  $\mu\text{g m}^{-3}$  at all times.**

487 \* AMS signal shown is OA + NO<sub>3</sub>, default

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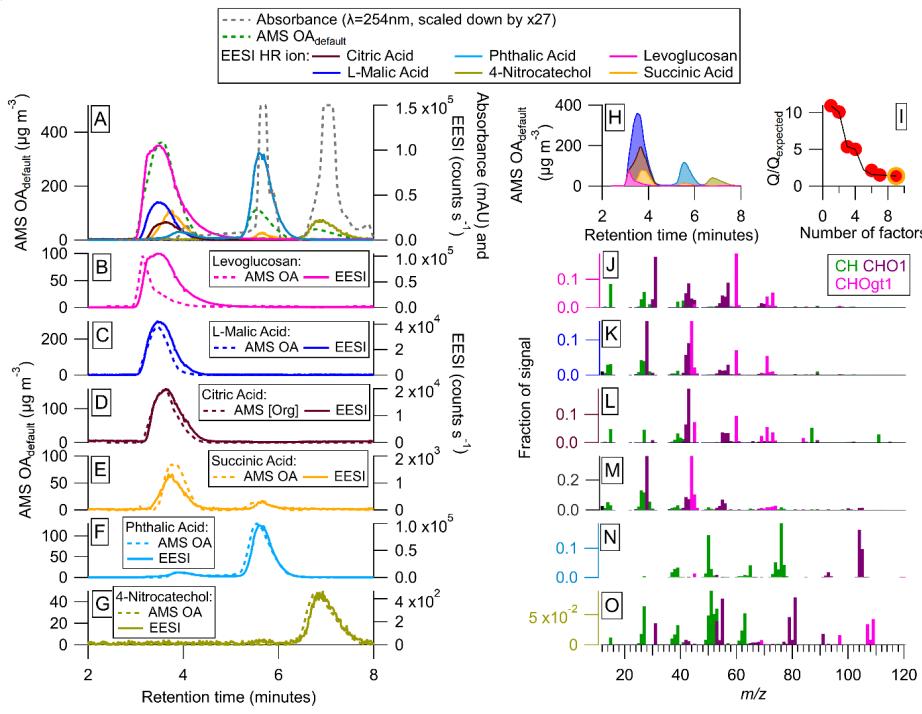
488 Figure 5a—Fig. 5f show excellent separation by PMF between the time series for each of the standards present in  
489 the mixture. This is likely due to the very different mass spectra for each species (Fig. 5h—Fig. 5l) as well as the  
490 time separation achieved by the HPLC. The mass spectra for each standard was compared to the direct calibration  
491 mass spectra to confirm the AMS PMF factors were assigned correctly (Fig. S6 and table S4). For all species, there  
492 was excellent correspondence, and the uncentered correlation coefficient (UC) between the mass spectral peaks was  
493 >0.95.

494 Here, the  $CF_x^A$  and  $CF_x^E$  values are known for each pure standard (from direct calibrations). When applying  
495 the  $CF$  to individual species, the overall agreement between the AMS and EESI time series is comparable to that

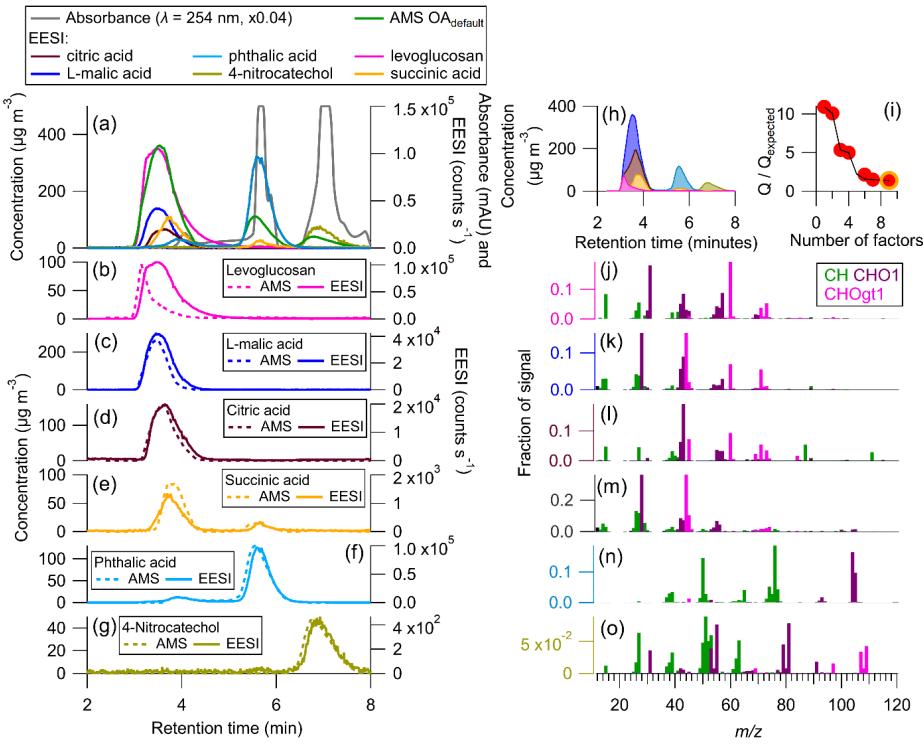
497 shown in Fig. 4. The AMS still underestimates succinic acid by a factor of  $\sim 2$  compared to the EESI, even after  
498 better separation is achieved with PMF. As discussed previously, this could be due to the mixing of the two species,  
499 which might change the viscosity or phase of the sampled aerosols compared to the pure species, which in turn  
500 could fundamentally change the  $CF_x^A$  due to the change in CE. Whilst separation was achieved with PMF, PMF time  
501 series are likely more accurate for systems where different species have similar  $CF_x^A$  (e.g. SOA mixtures from a  
502 single precursor and oxidant).

503 The [AMS chromatogram for the](#) mixture studied in Fig. 4 and Fig. 5 was mostly well-separated without  
504 PMF. In order to assess the ability of PMF to separate [AMS data for](#) a more complex mixture, PMF was run on a  
505 different standard solution shown in Fig. 6.





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508  
509 **Figure 6.** (Aa) time series of AMS total OA (assumed  $CF_x^A = 1.4$ ), EESI HR ion, and absorbance (max =  $4 \times 10^6$  milli-  
510 absorbance units). (Bb)-(Gg) AMS PMF factor (assumed  $CF_x^{A,\text{default}} = 1.4$ ) and EESI HR ion for 6 calibrants. (Hh)  
511 Stacked PMF factor solution time-series, (Gg)  $Q/Q_{\text{expected}}$  for AMS PMF solution, a 9-factor solution was chosen (yellow  
512 circle) with  $\text{FPEAK} = 0.2$ , and (Jj)-(Oo) AMS family-colored mass spectra for 6 PMF factors. **For levoglucosan and**  
513 **succinic acid, 2 factors were combined. The remaining factor was attributed to the background signal (< 2 µg m⁻³ at all**  
514 **times).**

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516 Unlike the data shown in Fig. 3 – Fig. 5, the species run in the standard solution shown in Fig. 6 were not calibrated  
 517 directly. Thus, Fig. 6 serves as a test of PMFs ability to resolve AMS data for complex mixtures, rather than a  
 518 comparison of the calibration methods. Figure 6a shows the uncalibrated time series/chromatogram for the  
 519 standards in the mixture. In contrast to the previous mixture, this solution contains 5-five co-eluting peaks:  
 520 levoglucosan, L-malic acid, citric acid, succinic acid, and a small fraction of the phthalic acid and its isomer. These  
 521 5-five co-eluting peaks suggest that the application of only HPLC with the separation method being used here is not  
 522 sufficient for these species, likely due to how polar they are. Further separation could be achieved by either  
 523 changing the HPLC method (through the use of a normal phase chromatography, which uses e.g. a silica column) or  
 524 running PMF on the AMS data.

525 Figures 6b – Fig. 6h shows AMS PMF time series for the standards present in the mixture. In Fig. 6b, both  
526 the AMS and EESI levoglucosan peaks have different shapes. The EESI peak has a right tail, which is potentially  
527 due to the “sticky” (semi-volatile) nature of levoglucosan (Brown et al., 2021). The AMS peak has a sharp increase  
528 and slow descent, and does not resemble a Gaussian (which is the approximate shape we expect eluting peaks to  
529 have). This is likely due to an imperfect PMF separation. Despite that, when comparing the mass spectra in Fig. 6j to  
530 the direct calibration mass spectra in Fig. S7, UC (table S5) is 0.93, suggesting consistency between the two mass  
531 spectra.

532 L-malic acid and citric acid also co-elute with levoglucosan. For citric acid, L-malic acid, and  
533 levoglucosan The PMF factors assigned to those species do look like Gaussian curves, but the mass spectra shown in  
534 Fig. 6j – Fig. 6l are somewhat similar. For L-malic acid and levoglucosan, m/z 60 makes up some of the observed  
535 signal. While m/z 60 is a known levoglucosan AMS ion, the direct calibration mass spectra for L-malic acid also  
536 shows some signal at m/z 60. The PMF mass spectra for L-malic acid has a slightly higher ratio of m/z 60 relative to  
537 the other ions, which could suggest that there is some mixing between the L-malic acid and levoglucosan factors.  
538 The assigned L-malic acid factor has a UC of 0.89 with the directly calibrated mass spectra, but citric acid was not  
539 directly calibrated for, and it is likely there is some overlap in the AMS factors between those three species. This  
540 was an especially complex solution for PMF to resolve due to the very similar retention times and mass spectra  
541 between these species.

542 As in Fig. 5, succinic acid, phthalic acid, and 4-nitrocatechol (Fig. 6e – Fig. 6g and Fig. 6m – Fig. 6o) are  
543 easily resolved when running PMF on the AMS HPLC-chromatograms. This is likely due to both the retention time  
544 differences and the different AMS mass spectra for these three species. In Table 1, calibration factors are shown for  
545 levoglucosan, succinic acid, phthalic acid, and 4-nitrocatechol.  $CF_x^A$  is known from the direct calibrations done in  
546 Fig. 4. During this experiment, only levoglucosan was cross-calibrated with a direct calibration, however, but the  
547 multi-instrumental calibration value is highly affected by the shape of the AMS PMF factor associated with  
548 levoglucosan. Thus, the multi-instrumental calibration factor for levoglucosan is likely incorrect. The PMF factor  
549 stacked time series is shown in Fig. 6h. These results suggest that while PMF run on the AMS data does provide  
550 further peak resolution compared to HPLC alone, PMF cannot completely resolve all co-eluting peaks.

### 551 3.3 Combined application of the multi-instrumental calibration method and PMF on $\beta$ -pinene + $\text{NO}_3$ SOA

552 In order to test the applicability of the proposed method to a complex real system, SOA from  $\beta$ -pinene +  $\text{NO}_3$  was  
553 generated, collected on a filter, extracted, and analyzed with our multi-instrument system (per Sect. 2.1). This SOA  
554 system has been studied in depth previously and 95 % of the SOA mass is composed of eight unique products,  
555 shown in Table 1 in Claflin and Ziemann (2018) and Table S6 here (Claflin and Ziemann, 2018). Of the eight  
556 known products, we identified molecular ions that are attributed to a monomer (m/z 268.1, assumed to be  
557  $[\text{C}_{10}\text{H}_{15}\text{NO}_4\text{-Na}]^+$  and five dimers. Some of the dimers elute as different isomers, but the EESI HR ions observed  
558 corresponded to m/z 451.2 ( $[\text{C}_{20}\text{H}_{32}\text{N}_2\text{O}_8\text{-Na}]^+$ ), m/z 467.2 ( $[\text{C}_{20}\text{H}_{32}\text{N}_2\text{O}_9\text{-Na}]^+$ ), m/z 483.2 ( $[\text{C}_{20}\text{H}_{32}\text{N}_2\text{O}_{10}\text{-Na}]^+$ ), and  
559 m/z 499.2 ( $[\text{C}_{21}\text{H}_{36}\text{N}_2\text{O}_{10}\text{-Na}]^+$ ), all of which were identified in Claflin and Ziemann (2018). We also observed two  
560 additional ions, m/z 388.2 and m/z 465.2, whose structures remain unknown. To better compare the differences in

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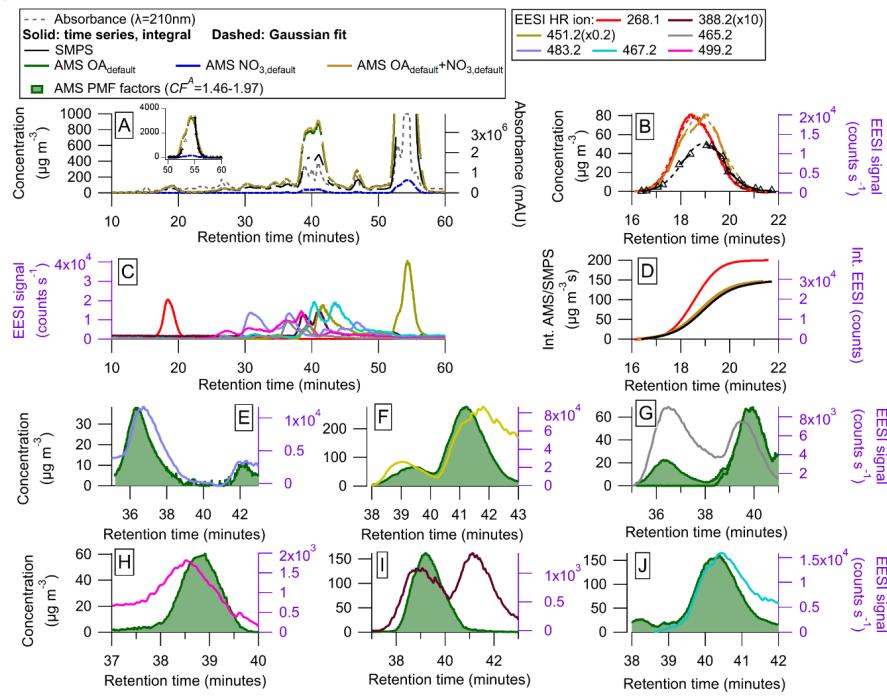
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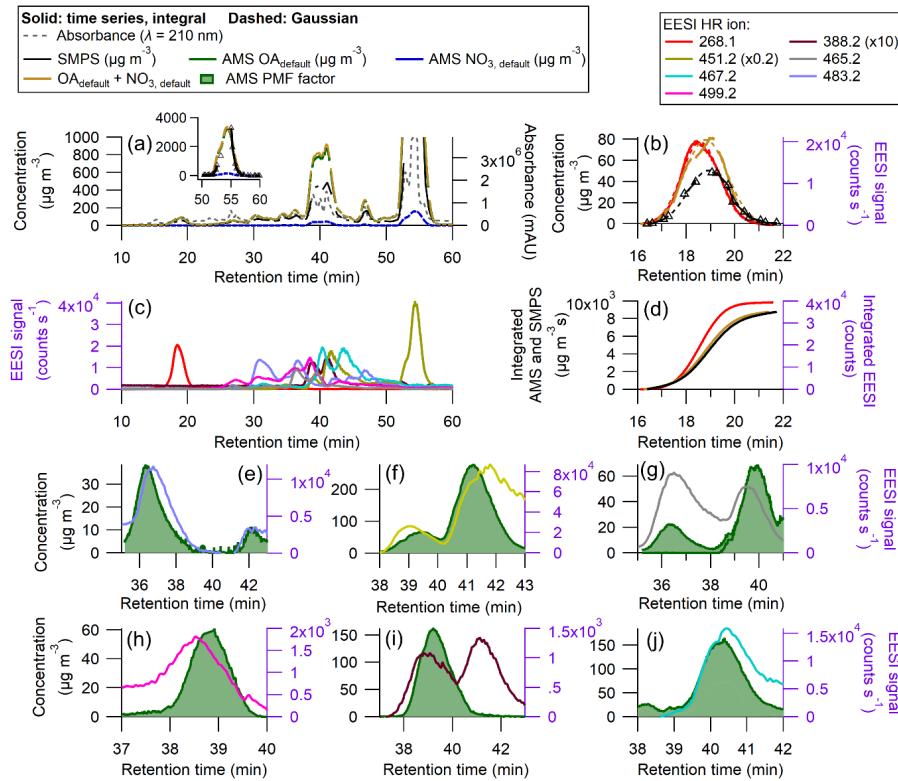
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561 the chromatogram obtained here vs that shown in Claflin and Ziemann (2018), we compare the UV-Vis time series  
562 in Fig. S9. The chromatograms are similar, although their chromatogram had slightly better resolution. Differences  
563 in observed species could potentially arise due to the age of the SOA extract used here (~ 1 year) vs. the fresh SOA  
564 extract used in that study, fragmentation of species in the EESI (e.g.  $m/z$  388.2), or other experimental factors. many  
565 of the products have been identified (Claflin and Ziemann, 2018; DeVault et al., 2022). The HPLC method was that  
566 of DeVault et al. (2022). Species here are identified based on comparison to the results in the aforementioned  
567 papers, and the observed EESI+ HR ions that show peaks in the time series (Fig. 7). Per Claflin and Ziemann  
568 (2018), many of the known products are oligomers, formed primarily from the reactions of two carbonyl nitrate  
569 monomers. For simplicity, the SOA peaks observed will be referenced by their associated EESI HR ion $m/z$ .

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571 **Figure 7. Results of an HPLC run for SOA from  $\beta$ -pinene + NO<sub>3</sub> (Aa)** AMS, SMPS, and UV-Vis chromatograms (milli-  
572 absorbance units), with inset showing peak from 50–60 minutes. (Bb) Time series and Gaussian fits for the peak between  
573 16 and 20 minutes (without using PMF). (Cc) EESI HR ions time series (Dd) time integrated mass concentrations (ion  
574 signal) for AMS OA and NO<sub>3</sub>, SMPS total mass, and EESI+ HR ion ( $m/z = 268.1$ ). (Ee)–(Jj) show some AMS PMF factors  
575 against measured EESI+ HR ions. (Gg), (Hh), and (Jj) represent split AMS PMF factors for the measured EESI+ HR ions.  
576 The AMS PMF factors have a  $CF_x^A$  ranging from 1.46–1.97 as shown in Fig. S3 and Table 2. Densities are applied to the  
577 SMPS data, shown in Fig. S8.

578 Figure 7a shows the full time-series for the  $\beta$ -pinene system. Many chromatographic peaks are observed by the  
579 AMS, SMPS, EESI, and UV-Vis. Many of the peaks are present in clusters and not well enough resolved enough to  
580 fit individual allow for the direct calculation of  $CF_x^A$  and  $CF_x^E$  using the SMPS as a the reference, as discussed in  
581 Sect. 2.7. Gaussian curves to the EESI and AMS data. Claijin and Ziemann (2018) measured a similar (albeit  
582 slightly better separated) UV-Vis chromatogram (Fig. S9). Differences could potentially arise due to the age of the  
583 SOA extract used here (~1 year) vs. the fresh SOA extract used in that study, or other experimental factors.

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585 The degree of peak co-elution is shown Overlapping peaks are also observed in the EESI data (in Fig. 7c). There are ←  
 586 two isolated peaks, from  $m/z$  268.1 from 15 - 21 minutes and  $m/z$  451.2 from 52 - 58 minutes. One peak, measured at  
 587 EESI HR ion  $m/z$  483.2 (suspected structure shown in table S6), was mostly resolved, and also shows up from 46 -  
 588 48 minutes. The raw (and fitted) data is shown in Fig. 7b for the EESI ion measured at  $m/z$  268.1 (a monomer,  
 589 triarboxylic nitrate) (Claflin and Ziemann, 2018). The integrated fits are shown in Fig. 7d.

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590 The EESI sensitivities for the overlapping peaks Multiple peaks overlap from ~30 to ~50 minutes (based  
 591 on the EESI data shown in Fig. 7c), were calculated by referencing the observed EESI signal to the AMS PMF time  
 592 series. These peaks are likely all dimers, the species identified by Claflin and Ziemann (2018) and measured by the  
 593 EESI are shown in Table 5. Not every peak observed in Claflin and Ziemann (2018) was identified here, which is  
 594 likely due to lack of EESI sensitivity to some species and potential decomposition of SOA products (specifically for  
 595 the trimer identified in Claflin and Ziemann (2018)). In contrast, some EESI HR ions that do not correspond to  
 596 peaks identified in Claflin and Ziemann (2018) were detected here, but structures for those species are unknown.

597 In Fig. 7e – Fig. 7j, AMS PMF time series that increased during for the middle third of the run are shown  
 598 alongside EESI HR ions. The full PMF solution can be found in Fig. S10 – Fig. S12. AMS factors were matched  
 599 with EESI HR ions based on the retention time and general shape of the time series. For some peaks, the retention  
 600 times differ by up to 0.5 min. These peaks are assigned based on the similarity in time series between the EESI and  
 601 AMS. The complexity of this solution, as well as the similarities in the products' molecular structures, likely  
 602 hindered the ability of PMF to fully resolve each individual product. For many of the overlapping peaks, that  
 603 overlap the most in time, the magnitude of the individual AMS PMF factors separated during this time are  
 604 comparable to each other.

605  $CF_x^E$  and  $CF_x^A$  are given for each identified species in Table 2. Many of the identified species have  $CF_x^E$  in  
 606 the same range as levoglucosan, within a factor of 3.

607  
 608 Table 2. EESI HR ion,  $CF_x^E$  (counts  $s^{-1} \mu g^{-1} m^3$ ),  $CF_x^E / CF_{lev}^E$ , and  $CF_x^A$ , and associated PMF factor for the  $\beta$ -pinene +  $NO_3^-$   
 609 SOA mixture.  $CF_x^A CF_{lev}^E / CF_{lev}^E = 441.6$  counts  $s^{-1} \mu g^{-1} m^3$ .  $CF_x^E$  was calculated using the AMS PMF [OrgOA]  $\times 1.05$   
 610 (the average  $[NO_3^-]$  contribution was ~5%, Fig. S3).

EESI ion	$CF_x^E$ (counts $s^{-1} \mu g^{-1} m^3$ )	$CF_x^E / CF_{lev}^E$ (unitless)	$CF_x^A$ (unitless)	AMS PMF factor(s)
268.1	270	0.61	1.46	-
388.2	10.9	0.023	1.97	9, 13
451.2 (1)	407	0.92	1.97	13
451.2 (2)	423	0.96	1.73	13
451.2 (3)	83.2	0.19	1.97*	-
465.2 (1)	670	1.5	1.97	2
465.2 (2)	170	0.38	1.97	10

467.2	139	0.31	1.73	5,8
483.2	435	0.99	1.97	44
499.2	54.2	0.12	1.97	42

611 \* Incomplete SMPS data, assuming  $CF_x^A = 1.97$ .

EESI ion	$CF_x^E$ (counts $s^{-1} \mu g^{-1} m^3$ )	$CF_x^E / CF_{levo}^E$ (unitless)	$CF_x^A$ (unitless)
268.1	270	0.61	1.46
388.2	10.9	0.023	1.97
451.2 (1)	407	0.92	1.97
451.2 (2)	423	0.96	1.73
451.2 (3)	83.2	0.19	1.97*
465.2 (1)	670	1.5	1.97
465.2 (2)	170	0.38	1.97
467.2	139	0.31	1.73
483.2	435	0.99	1.97
499.2	54.2	0.12	1.97

612 \* Incomplete SMPS data, assuming  $CF_x^A = 1.97$ .

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614 Some species, like the EESI HR ions measured at  $m/z$  388.2 and  $m/z$  499.2, have much lower EESI sensitivity than  
615 the other species. These species could be fragments of a larger parent ion, or they could be species that, for whatever  
616 reason, do not form a strong adduct with  $Na^+$ . The ambiguity in the PMF factors may result in some errors in  $CF_x^E$ ,  
617 but they are unlikely to fully explain the factor of 10<sup>ten</sup> difference in sensitivity between the most and least sensitive  
618  $\beta$ -pinene +  $NO_3$  products. In future runs with slightly better chromatographic separation a multi-variate fit of  
619 individual factors vs. the SMPS may allow further constraining the quantification.

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621 In this system, many of the products differ only by one or two oxygen atoms. Some may contain a In some  
622 cases, a carboxylic acid functional group replaces in the place of a ketone, whilst others other molecules contain a  
623 cyclic ether, and some do not. The subtle differences in structure could influence the sensitivity with the EESI, as  
624 the oxygenated moieties may change the likelihood of forming a strong  $[M+Na]^+$  adduct. Further, some EESI HR  
625 ions eluted multiple times (e.g.  $m/z$  451.2). Claflin and Ziemann (2018) identified the structure of this ion for the  
626 third peak (shown in Table S6). However, this ion is measured twice more, from 38 - 43 minutes, which suggests the  
627 presence of isomers. Isomers can have different structures (shown in Table S6) and different  $CF_x^E$ . One example is  
628  $m/z$  483.2, where one isomer has a  $CF_x^E = 327.2$  and a second isomer has a  $CF_x^E = 54.2$  counts  $s^{-1} \mu g^{-1} m^3$ . Due to the  
chromatographic separation between these peaks and the third peak, it is likely that the first two species are some

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629 isomeric form of the species identified in Claflin and Ziemann (2018). As is shown for  $m/z$  483.2 (Table S6),  
630 isomers can have different structures and very different  $CF_x^E$  (327.2 vs. 54.2 counts  $s^{-1} \mu g^{-1} m^{-3}$ ).

631 Despite differences in  $CF_x^E$ ,  $CF_x^A$  was more consistent. In table 2, the AMS response to different SOA  
632 species formed from a single VOC precursor varies only by 25 %. For the mixed peaks (individual EESI  $m/z$ 's  
633 shown in Fig. 7e 7j),  $CF_x^A$  was either 1.48-97 or 1.5873, as shown discussed in Sect. S3 and shown in Fig. S3.  
634 For the three isolated peaks ( $m/z$  268.2,  $m/z$  451.2 [peak 3], and  $m/z$  483.2 [peak 2]), the  $CF_x^A$  spanned from 1.31 to  
635 1.75. For one of the isolated peaks,  $m/z$  451.2 (peak 3), the actual  $CF_x^A$  was not calculated, due to a malfunction of  
636 the SMPS system between 54-56 minutes. Individual peaks' Gaussian fits and integrated curves are shown in Fig.  
637 S13.

### 638 3.4 Discussion on the application of this method

639 In this paper, a novel technique was introduced that allows for the calibration of real-time mass spectrometers for  
640 individual species that cannot be obtained directly. This paper addresses the feasibility, performance, and limitations  
641 of this technique, all of which are necessary for any future use of this method.

642 The original purpose of this method was to calibrate species in SOA formed from laboratory chamber  
643 experiments. In many cases, the identity of the species was unknown, or the species could not be purchased as a pure  
644 standard. During those chamber experiments, SOA composition was measured in real-time with AMS, EESI, and  
645 SMPSs. SOA was also pulled through a Teflon filter, extracted in solvent, injected into the HPLC.

646 One application of this method would allow calculating yields for different SOA species produced from the  
647 oxidation of individual VOCs. This would allow for a better understanding of the chemical and partitioning  
648 mechanisms controlling the SOA composition and formation, along with providing information on which species are  
649 contributing the most to environmental and human health issues caused by SOA (e.g. higher light absorption or  
650 increased toxicity).

651 Another application is inferring calibration factors for important species in field datasets. This could be  
652 done by collecting filters to use with this method, including using UPLC for higher resolution. Alternatively, if  
653 specific primary sources or SOA precursors are known to be important for a dataset, those can be sampled in the lab  
654 to determine key species and their calibration factors.

655 One example of a field application is the FIREX-AQ field campaign, where the Jimenez lab at the Univ. of  
656 Colorado Boulder operated an EESI (Pagonis et al., 2021). During that campaign, direct calibrations were performed  
657 daily using either 4-nitrocatechol or levoglucosan. In the laboratory, these calibrations were also carried out daily,  
658 before chamber experiments and before running the HPLC calibration method. If species specific sensitivities are  
659 obtained in the lab, then they can be ratioed to either 4-nitrocatechol or levoglucosan, providing the relative  
660 sensitivity of individual analytes. The relative sensitivity can be referenced to the sensitivities obtained in the field,  
661 allowing for the budgeting of ambient SOA for multiple species.

## 662 **4 Conclusions**

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663 A-In this study, we introduced a novel multi-instrumental calibration method for EESI and AMS has been  
664 demonstrated here, that uses the HPLC and PMF to separate complex standard mixtures and SOA into individual  
665 species or sub groups of species present in the mixture. The chemical separation power of the HPLC, combined with  
666 analytical aerosol detection of SMPS, AMS, and EESI to calibrate the mass spectrometers for individual species in  
667 mixtures. When running Our proof of concept test using individual pure standards demonstrated close agreement  
668 (within 20 %) between direct and multi-instrumental calibration factors, indicating this method's quantitative ability.  
669 In a second proof of concept using a mostly resolved standard mixture, EESI direct and multi-instrumental  
670 calibration factors agree within a factor of two for low volatility species. We note that this method is not suitable for  
671 semivolatile species whose C\* is similar or higher than the concentration of aerosol sampled inside the SMPS DMA  
672 column. These results suggest that this method can be used to reliably determine species sensitivities for completely  
673 and mostly resolved chromatograms. In individual standards, the multi-instrumental calibration method agreed with  
674 direct calibration within 20%. As the sensitivities of EESI measured species can vary by over an order of magnitude,  
675 quantification within 20% is very useful. In a mixed standard run that contained mostly resolved species, the EESI  
676 and SMPS agreed within a factor of 1.5 (for non-volatile species). The AMS and EESI matched moderately well,  
677 except when measuring succinic acid.

678 When HPLC alone failed to fully resolve individual analytes, In situations where the HPLC column/method  
679 was unable to fully separate injected components, PMF on AMS data successfully resolved individual analytes time  
680 series in a simple standard mixture. However, in more complex standard and SOA mixtures, while PMF provided  
681 some additional chromatographic separation, the PMF solution showed signs of factor mixing. This was especially  
682 evident in the was used to methodically compare the time series and mass spectra for different species, and generate  
683 time resolved OA data for the AMS. This was especially important for the AMS data, as overlapping peaks are  
684 measured as large and wide “total OA” peaks for that instrument.

685 The  $\beta$ -pinene +  $\text{NO}_3$  SOA solution mixture, which contained many similar analytes, was the most complex mixture  
686 studied here, resulting in a less well resolved PMF solution, primarily due to the suspected presence of many  
687 isomers. The majority of the SOA peaks overlapped during the middle third of the HPLC run. PMF separation  
688 While approximate EESI and AMS calibration factors were obtained, these sensitivities are affected by the inherent  
689 error in the PMF solution, conducted on the HPLC separate AMS results produced a more complicated solution than  
690 the AMS PMF done on the standards' runs. This was likely due to similarities in mass spectra and retention times  
691 for the overlapping peaks. Despite that analytical challenge, when the middle third of the chromatogram was  
692 scrutinized using both the AMS PMF solution and the measured EESI+ HR ions, approximate calibration factors  
693 were obtained. In practice, while some mixtures may be adequately resolved by HPLC alone, AMS PMF can  
694 improve the chemical resolution of complex systems.

695 Future studies should prioritize improving the chromatography for the system of interest, potentially  
696 through changing the column type and / or mobile phase gradients, or using systems with higher intrinsic resolution  
697 such as UPLC (Kenseth et al., 2023). For future studies, additional effort should be focused on tuning the HPLC  
698 performance (e.g. through changing the column or mobile phase gradients) that provides higher resolution for  
699 whatever system is being studied. In this demonstration project During the experiments shown in this manuscript we

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700 were limited to a C<sub>18</sub> column, ~~which is most often used~~which is primarily suited for ~~separating~~ less polar species. ~~In~~  
701 ~~many situations, especially when there is~~However, in the polar standard mixtures shown here and in scenarios  
702 involving significant oxidation and smaller precursor gases, the resulting products are likely ~~too be more polar~~ to be  
703 ~~adequately than can be separated~~ by a C<sub>18</sub> column. In those experiments, a column with a polar stationary phase  
704 would allow for the separation of SOA components. In future experiments, columns with more polar stationary  
705 phases should be considered. If HPLC separation alone could completely resolve all chemical peaks, then PMF  
706 would not be needed, however in practice it is likely to help the chemical resolution of complex systems.

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707 In conclusion, our method offers a valuable tool for quantifying EESI and AMS sensitivities in mixtures,  
708 especially pertinent for laboratory generated SOA lacking pure standards or characterized by unknown isomeric  
709 forms. This technique can also be applied to other real-time aerosol mass spectrometers. To our knowledge, this  
710 technique stands as one of very few available methods for rapid calibration of EESI and AMS for SOA species that  
711 are unavailable as pure standards, emphasizing its significance in atmospheric research. These results introduce a  
712 new technique for better quantifying the instrument responses of the EESI and AMS to different molecular species  
713 present in complex mixtures such as from biomass burning, urban, and/or biogenic SOA.

## 714 5 Acknowledgements

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718 CIRES graduate research fellowship.

## 719 6 Author Contributions

720 MKS, DAD, JLJ, and PJZ designed the experiments, MKS carried them out with support from DAD, DK, SY, and  
721 PCJ. ACZ, PJZ, and MPD provided the HPLC instrument support. MKS carried out all data analysis and preparation  
722 of the manuscript, with contributions from all coauthors.

## 723 7 Competing Interests

724 The authors declare that they have no conflict of interest.

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981 5637–5653, doi:10.5194/amt-9-5637-2016, 2016.  
982

983

1 **Supporting information for:**

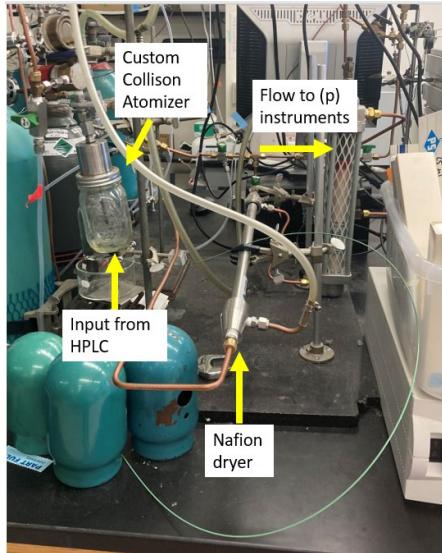
2 **“A multi-instrumental approach for calibrating two real-time**  
3 **mass spectrometers using high performance liquid**  
4 **chromatography and positive matrix factorization”**

5 **Melinda K. Schueneman<sup>†</sup>, Douglas A. Day<sup>†</sup>, Dongwook Kim<sup>†</sup>, Pedro Campuzano Jost<sup>†</sup>, Seonsik**  
6 **Yun<sup>†</sup>, Marla P. DeVault<sup>†</sup>, Anna C. Ziola<sup>†</sup>, Paul J. Ziemann<sup>†</sup>, and Jose L. Jimenez<sup>†</sup>**

7 <sup>†</sup>Department of Chemistry and Cooperative Institute for Research in Environmental Sciences, University of  
8 Colorado, Boulder, CO 80309, USA

9

10 S1 General system information for multi-instrumental calibration method



11

12 **Figure S1. HPLC tubing into custom atomizer**

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14 **Table S1. Tube volumes, flows, and residence times from HPLC separation to particle instrument detection.**

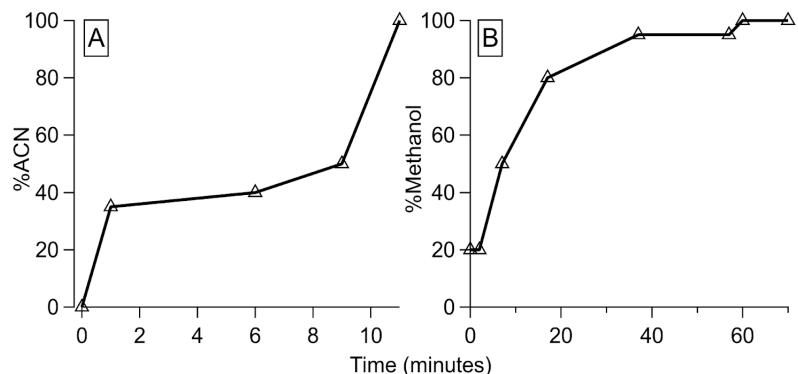
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Item	Total volume (mL)	Flow rate (flow through)	Residence time
Tubing transferring liquid from after HPLC column and UV-Vis detection to atomizer	0.67	1.0 mL min <sup>-1</sup>	40 s
Atomizer	500	8.0-10.4 mL min <sup>-1</sup>	3.0 - 3.758 s
Nafion drier	7.0	~8.04 mL min <sup>-1</sup>	0.053 s
Tubing before manifold	14208.3	7.24 mL min <sup>-1</sup>	1.0 s
Post manifold EESI	31.2	0.844 mL min <sup>-1</sup>	2.2 s
Post manifold AMS	14.4	1.54 mL min <sup>-1</sup>	0.60 s

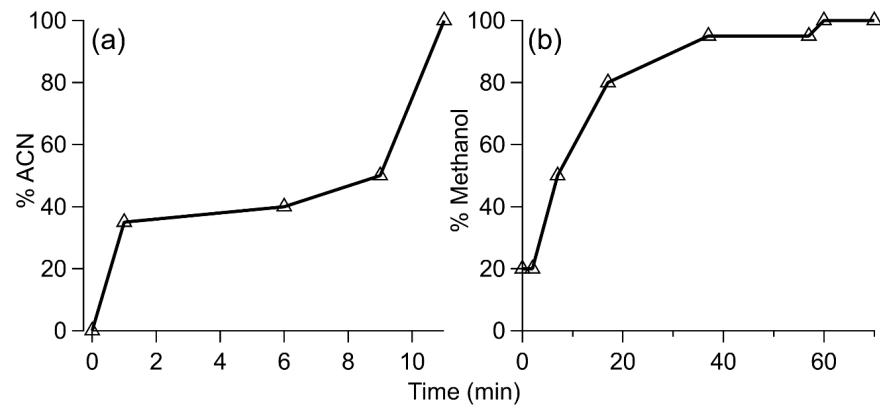
Post manifold SMPS A	34.2	1.43 <u>HL</u> min <sup>-1</sup>	1.4 s
Post manifold SMPS B	28.59	1.49 <u>HL</u> min <sup>-1</sup>	1.2 s

15

16



17



18

19 **Figure S2.** Solvent gradients for (Aa) standard HPLC runs and (Bb)  $\beta$ -pinene HPLC run. The other solvent  
20 was a mixture of 95 %  $\text{H}_2\text{O}$  / 5 % ACN.

21

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22 **Table S2. Standard compounds used for HPLC method demonstration, source and purity, volatility**  
23 (calculated using published vapor pressures), estimated percent evaporated during transmission (from  
24 atomizer output to detection, calculated with  $C_{\text{ea}}^*$  and measured OA concentration at detection), and density  
25 (using the ratio of  $d_{\text{ea}}/d_m$ )

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Species	Source + purity	Saturation Mass Concentration ( $\mu\text{g m}^{-3}$ ) (T=298 K)	Estimated Percent Evaporated	Density
3-methyl-4-nitrophenol	Aldrich, 98 %	5,210	92 %	1.27**
Phthalic acid	Beantown Chemical, ACS grade, 99.5 %	5.72	0 %	1.05
4-nitrophenol	Aldrich, 99 %	10,600	94 %	1.48**
Succinic acid	Aldrich, 99 %	1.21	0 %	1.18
4-nitrocatechol	Alfa Aesar, 98 %	64	63 %	1.26
L-malic acid	Aldrich, 97 %	0.24	-	1.28
Citric acid	Fisher Scientific	0.18	-	-
Levoglucosan	Chem-Impex Int'l, $\geq 99.0\%$	13*	-	1.30
Acetonitrile	Fisher Chemical, $> 99.95\%$	-	-	-
Methanol	Fisher Chemical, $> 99.9\%$	-	-	-
Water	VWR Chemicals, HPLC grade	-	-	-
Ethyl Acetate	Fisher Chemical, 99.5 %	-	-	-

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26 \*Reported in (Pagonis et al., 2021)

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27 \*\*Density of bulk solution from literature

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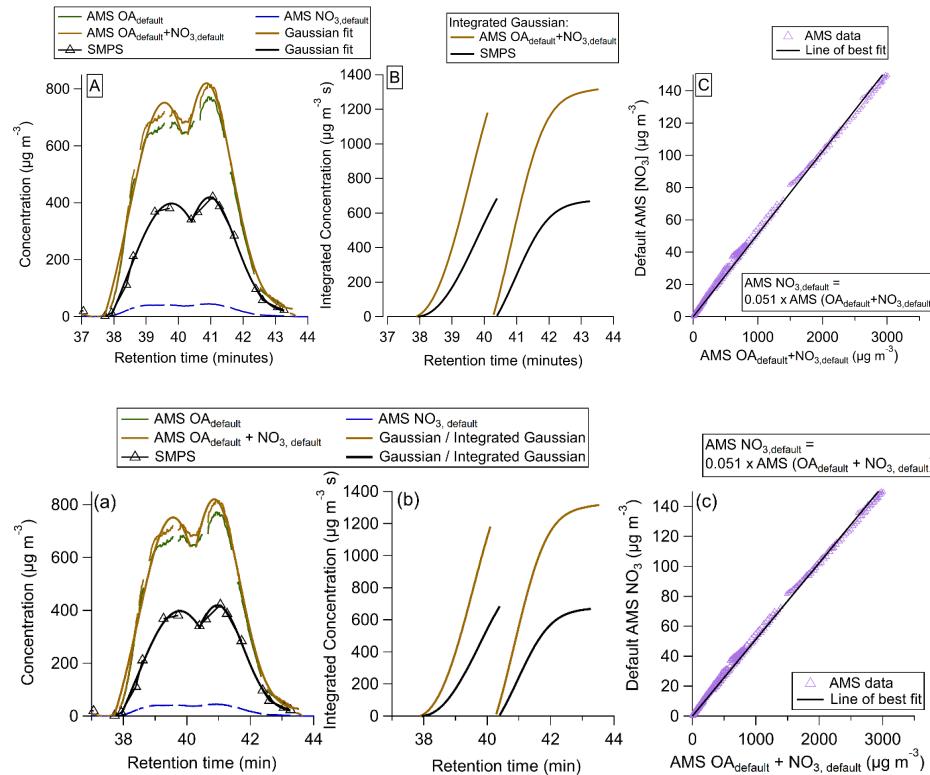
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28 The densities measured using the  $d_{va}/d_m$  ratio do not match the literature values for bulk density well. This is  
 29 potentially due to different phases from that of the bulk material, and/or non-spherical particle shape (Jayne et al.,  
 30 2000; Huffman et al., 2005). Regardless, the  $d_{va}/d_m$  density was used as the best estimate here.

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32  
 33 **Figure S3.** (Aa) AMS default mass concentrations for  $[OA]$ ,  $[NO_3]$ , and  $[OA + NO_3]$ ; SMPS mass  
 34 concentrations, corrected for the average density. (Bb) Integrated Gaussian curves for each peak. (Cc)  
 35 Default AMS  $[NO_3]$  vs total default AMS signal  $[OA + NO_3]$ , fit with a line. The slope (ratio of  $[NO_3]/[OA +$   
 36  $NO_3]$ ) = 0.051.

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37  
 38 The nitrate contribution to the total mass for this peak was ~5.1 %. Fitting the bulk peaks (which are composed of  
 39 multiple eluents) may result in some error in the nitrate contribution approximation.  $CF_x^A$  was calculated for the two

41 peaks by referencing the AMS mass to the SMPS mass. For the first peak,  $CF_x^A = 1.97$ , for the second peak  $CF_x^A =$   
42 1.73.

43 **S2 SMPS testing and validation**

44 **S2.1 Fast scanning operation and validation**

45 The fast scanning operation of the SMPSs was essential here. A “fast scan” here means 30 ~~seconds~~ for voltage  
46 scanning, with 10 ~~seconds~~ retrace time (when the voltage is returned back to 0). This allows for an SMPS data point  
47 to be obtained every 40 ~~seconds~~, and when two SMPSs are used with interleaved timing, every  $\sim$ 20 ~~seconds~~. This  
48 faster scanning is not without precedent; one paper published in 1990 first denoted the term “scanning electrical  
49 mobility spectrometer” or SEMS (Wang and Flagan, 1990). In that paper, researchers demonstrated that aerosol  
50 distributions for atmospherically relevant samples could be measured in a 30-~~second~~ scan time, with a 30-~~second~~  
51 retrace time. This research led to the creation of new SMPSs that, like the SEMS, scanned continuously, and thus  
52 would be capable of 30  $s$  scanning times. A study a few years later put this to the test, and looked at the impact of  
53 changing SMPS scan times, and found that shorter scan times led to more smearing (less-resolved size distributions)  
54 and lower peak maximas (Russell et al., 1995). They suggest that this is driven by the residence time of the particles  
55 from the output of the DMA to the optical detection by the CPC ( $t_d$ ). In addition, a paper in 2002 elaborated on the  
56 conclusions from Russell et. al. (1995), and found that when scanning with a flow rate of 0.3  $L\text{ min}^{-1}$ , combined  
57 with a 30 ~~seconds~~ scan time, the size distribution was significantly broadened (Collins et al., 2002). The maximum  
58 concentration was decreased by over 50 % when compared to a longer scan time (300  $s$ ), but the integrated  
59 concentration did not seem as affected, due to broadening in the faster scan.

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60 Typically, SMPSs are run at longer scan times of ~~two-2 minutes~~ or more (Sioutas, 1999; McMurry, 2000;  
61 Jeong and Evans, 2009). One study modified an SMPS by adding an aerosol particle mass analyzer (APM). With the  
62 modified system, data points were recorded every 60 ~~seconds~~ (Malloy et al., 2009). Another study, which took place  
63 on an aircraft and measured the air over Mexico City, ran their SMPS with a scan time of 1.5 ~~minutes~~ (DeCarlo et  
64 al., 2008). Despite the conclusions of Wang and Flagan (1990), many in the community run their SMPSs as “slow”  
65 (e.g. scan times of two or more ~~minutes~~) instruments. Henceforth, “slow” will refer to the 2 ~~minute~~ scans, and “fast”  
66 will refer to the 30 ~~second~~ scans.

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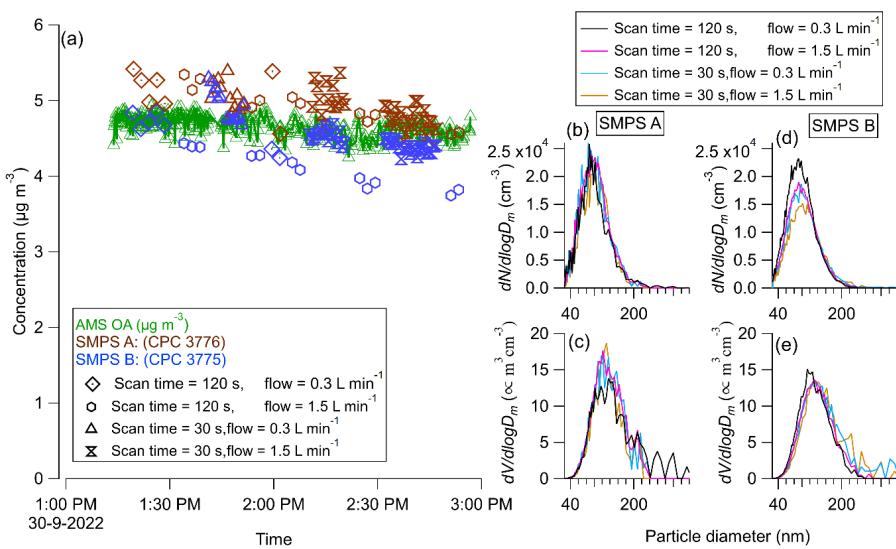
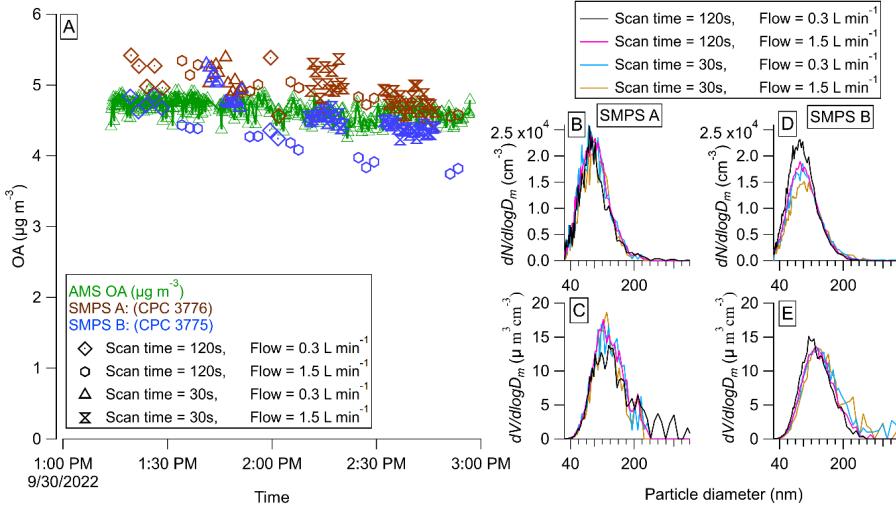
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67 Here, we test each SMPS with a combination of “long” scans (2-~~minute~~ scans, 15 s retrace, 3  $L\text{ min}^{-1}$   
68 sheath flow) and “fast” scans (30  $s$  scans, 10 s retrace, 6  $L\text{ min}^{-1}$  sheath flow). In order to assess the usability and  
69 accuracy of the fast scan method, tests were carried out (Fig. S4) to compare the total integrated volume  
70 concentration, number size distributions, and volume size distributions for ~~two-minute min~~ scans at both a sample  
71 flow of 0.3  $L\text{ min}^{-1}$  and 1.5  $L\text{ min}^{-1}$ , and 30 ~~second~~ scans done with the same flow rates.

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75 **Figure S4. (Aa)** Estimated particle mass concentration from SMPS A and B compared to the total OA  
76 measured by the AMS, for different combinations of scanning times and sample flow rates when sampling  
77 constant DOS concentrations from a large chamber. **(Bb)** Number distribution comparisons for different  
78 combinations of scanning times and flow rates for SMPS A, **(Cc)** Volume distribution comparisons, **(Dd)**  
79 number distribution comparisons for SMPS B, and **(Ee)** volume distribution comparisons for SMPS B.

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81 In Fig. S4a, the total concentration of dioctyl sebacate (DOS) was measured by an AMS (green) and time averaged  
 82 to 10 ~~seconds~~. The AMS-~~measured~~ DOS (after AMS calibration for that species) was used as the reference  
 83 concentration. DOS was generated using a custom evaporation-condensation apparatus (Sinclair and La Mer, 1949;  
 84 Krechmer et al., 2017) (Sinclair and La Mer, 1949), and flowed into a 20 m<sup>3</sup> Teflon chamber. To start, we scanned  
 85 with both SMPSs set to a 2 minute scan time with a 15 ~~second~~ retrace time, and a flow rate of 0.3 L min<sup>-1</sup>. This is  
 86 typically how we run our SMPSs for laboratory studies and we have compared with even longer scans (up to 300 s,  
 87 same flow settings) showing good agreement (Liu et al., 2019) and has shown good quantitative agreement for  
 88 intercomparisons during chamber and field campaigns. Those “long scans” serve as a reference. Both SMPSs were  
 89 run concurrently.

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90 Some researchers show peak smearing when using faster scan times (although, those studies seem to use a  
 91 sample flow rate=0.3 L min<sup>-1</sup>) (Russell et al., 1995). These studies posit that the smearing is mainly due to  
 92 instrument specific/plumbing delay times from the output of the DMA to the optical detection by the CPC (Russell  
 93 et al., 1995). In Fig. S4b, the number distribution is shown for the different flow/scan time configurations for the  
 94 SMPS A. The black distribution for all scans is the reference (120 s scan, 0.3 L min<sup>-1</sup>, resolution=10). For the  
 95 number distribution, the peak width for the reference is more narrow than for all other configurations. The difference  
 96 is minor, however, and not as large as in other reports.

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97 In Fig. S4c, the volume distributions are compared. The reference scan has a lower maximum concentration  
 98 than the other configurations, which seems to go against previously published results. Over time, [DOS] measured  
 99 by the AMS decreases, due to chamber wall loss effects. To counter this, reference scans (120 scans, 0.3 L min<sup>-1</sup>  
 100 flows) are carried out throughout the experiment. For reference, the SMPSs were run with 30 s scans and 1.5 L  
 101 min<sup>-1</sup> sample flows for the HPLC method proposed in the main text.

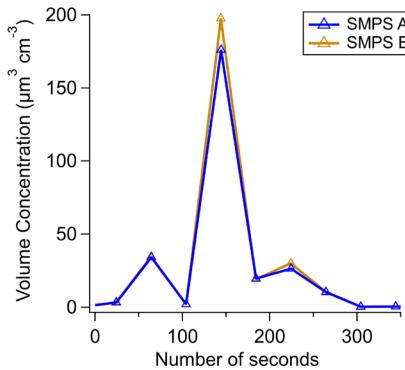
102 The distributions for SMPS B are more affected by the different configurations. This is unsurprising, as it  
 103 has a longer  $t_d$  than SMPS A (table S3), and likely is more representative of the systems studied in the research cited  
 104 above. In Fig. S4d, the number distribution for the reference scan has a ~~higher maximum~~ ~~higher maximum~~ than the  
 105 other scans. The faster, high flow scan is the most different from the reference, and has both a lower maxima and a  
 106 wider peak width (resolution = 4). This matches previous findings (Collins et al., 2002), but this study shows a far  
 107 less dramatic peak shape difference than that shown therein. This finding could introduce some quantification error.  
 108 In Fig. S4, the volume distributions match fairly well for all configurations. A faster instrument (such as an optical  
 109 particle counter) would be ideal to obtain faster measurements, but the small diameter particles produced by the  
 110 Collison atomizer makes running those instruments impractical and prone to error (due to low detection efficiency at  
 111 smaller size particles).

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112 For the multi-instrumental calibration experiments, SMPS A and SMPS B were offset by ~~twenty~~ 20  
 113 ~~seconds~~. That allowed us to obtain a volume concentration every approx. 20 ~~seconds~~. For comparing the response  
 114 between the two SMPSs, an experiment was done where SMPS A and SMPS B were run concurrently (Fig. S5).

115 SMPS A and SMPS B are shown to match within  $\sim 0\% \pm 10\%$  (at the maxima). The consistency observed in Fig.  
116 S5 between SMPS A and SMPS B provides increased confidence in the use of each instrument in “fast” mode.

117



118

119 **Figure S5. Concurrent SMPS scans for an HPLC run**

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## 120 **S2.2 SMPS delay time calculations**

121 Delay times from the aerosol sampling manifold to the DMAs were calculated by running each DMA to size select  
122 particles with a mobility diameter of 115 nm. Following transmission, the time it takes for the CPC concentration to  
123 reach half of its maximum concentration ( $t_{1/2}$ ) was calculated (table S3). Here, delay times were short, due to the  
124 high sample flow. This does not eliminate the importance of having accurate delay times. Fast scans are often prone  
125 to more error than their slow counterparts.

126 To calculate  $t_d$  (table S3), polystyrene latex spheres (PSLs) of a known diameter were atomized and  
127 measured by the SMPSs. Calculating delay times ( $t_{1/2}$  and  $t_d$  [delay time from exit of the DMA to the CPC]) allowed  
128 us to properly align the slower SMPS measurements with the fast mass spectrometer measurements during the  
129 relatively short chromatographically-separated compound peaks. Each eluting HPLC peak is only approx. 1.5  
130 minutes long, and the instruments are run at different time resolutions. Each SMPS collects one data point every 40  
131 seconds. For each data point, the SMPS software provided an uncorrected scanning start time. During the 40-second  
132 scan, concentrations can change significantly. If the SMPS scan starts 15 seconds before the maxima is reached,  
133 then the scan is recording concentrations at particle diameters both before, during, and after the peak maxima. If the  
134 SMPSs were not corrected for their delay times, then the SMPS data point would show an erroneously low/high  
135 concentration, and lead to errors when comparing to the other instruments.

136 **Table S3. Delay times for each SMPS.  $t_{1/2}$  is the time it takes for the CPC concentration to reach half of its  
137 maximum concentration**

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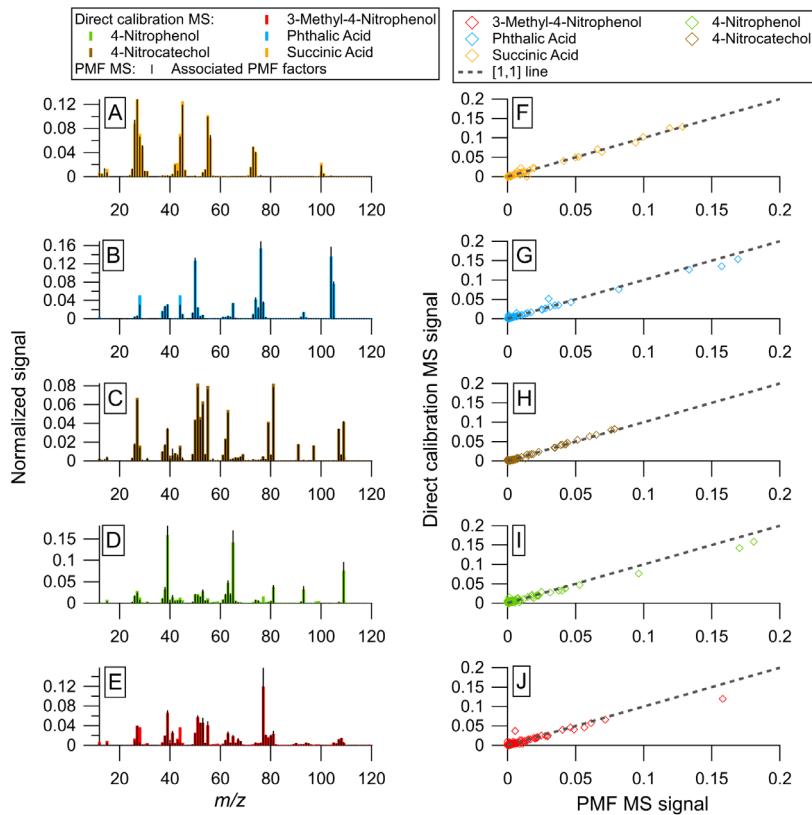
SMPS name	CPC type	Delay time ( $t_{1/2}$ ) (s)	DMA $\Rightarrow$ to CPC delay time ( $t_d$ ) (s)
SMPS A	3776	10.5	0.43
SMPS B	3775	8	1.55

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139 **S3 Standard mixture mass spectra comparison for direct and multi-instrumental calibrations factors**

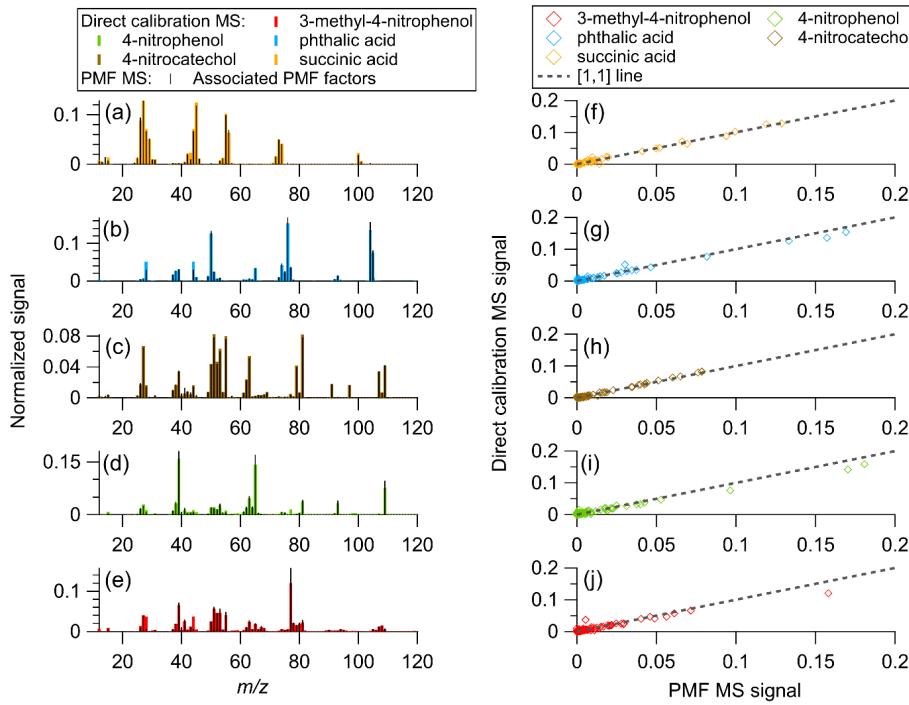
140 Mass spectra were obtained from PMF for many of the standards used in Sect. 3.2 and compared against the average  
141 mass spectra from direct calibrations (Fig. S6).

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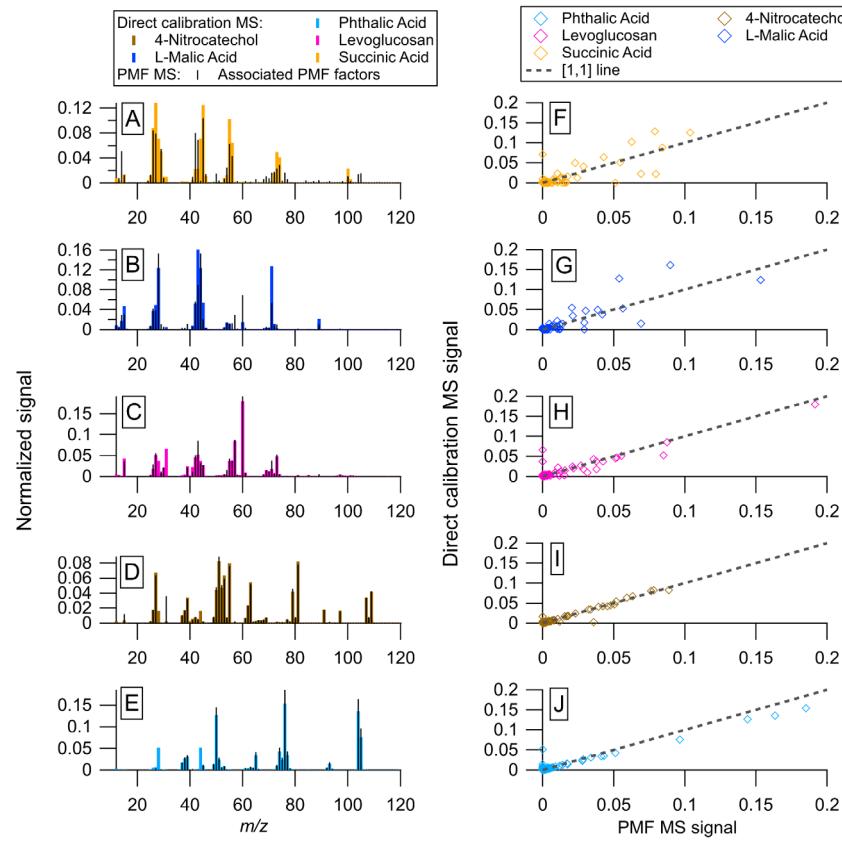
PMF factor MS	Succinic aAcid	<u>4-</u> <u>Nitrocatechol</u> <u>nitrocatechol</u>	Phthalic Acidacid	<u>4-</u> <u>Nitrophenol</u> <u>nitrophenol</u>	3-methyl-4- nitrophenol
Succinic <u>Acidacid</u>	<b>0.99</b>	0.38	0.14	0.15	0.30
<u>Nitrocatechol</u> <u>4-</u> <u>nitrocatechol</u>	0.38	<b>1.0</b>	0.23	0.49	0.62
Phthalic <u>aAcid</u>	0.094	0.20	<b>0.99</b>	0.24	0.31
<u>Nitrophenol</u> <u>5-nitrophenol</u>	0.10	0.43	0.24	<b>0.99</b>	0.45
3-methyl-4- nitrophenol	0.21	0.58	0.27	0.49	<b>0.96</b>

154

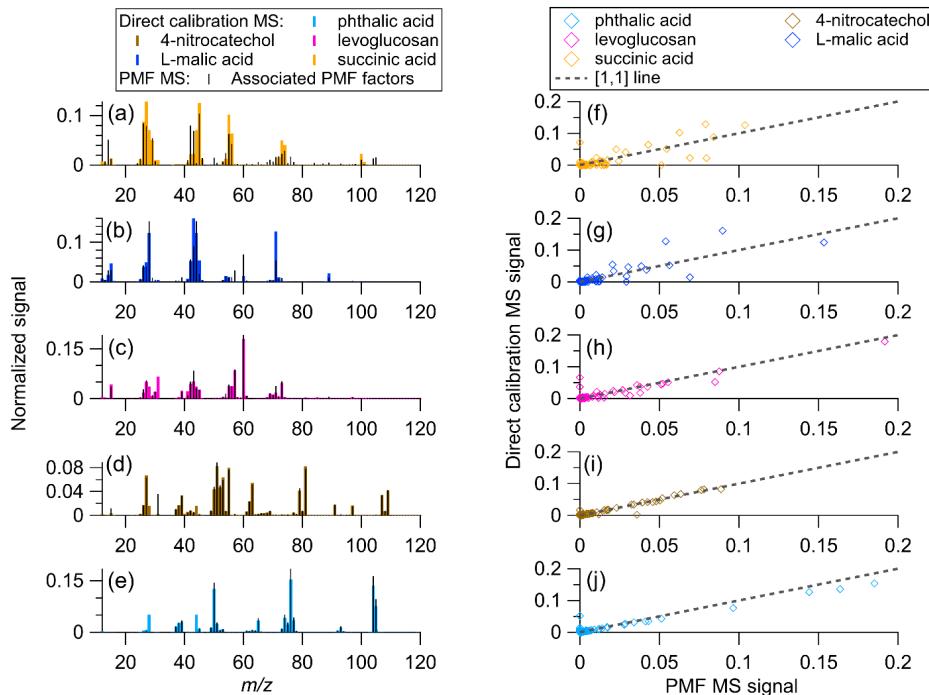
155 The UC provides the same information as the dot product, without the need to normalize the mass spectra. For all  
156 species, the UC  $> 0.95$ . For 4-nitrocatechol, the UC rounded up to 1.0 (near perfect agreement).

157 Similarly to the process carried out above, the mass spectra from the PMF solution for the data shown in  
158 Fig. 6 was compared to direct calibrations (Fig. S7).

159



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169

**Table S5. Uncentered correlation coefficient (UC) between AMS direct calibration and PMF factor mass spectra (Ulbrich et al., 2009) for standard solution 2 (Fig. 6, Fig. S7)**

	Direct calibration MS				
PMF factor MS	Succinic Acid	Malic Acid	Levoglucosan	4-Nitrocatechol	Phthalic Acid
Succinic Acid PMF	<b>0.81</b>	0.50	0.35	0.31	0.17
Malic Acid PMF	0.55	<b>0.89</b>	0.60	0.20	0.23
Levoglucosan PMF	0.36	0.41	<b>0.93</b>	0.19	0.029
Nitrocatechol-4-nitrocatechol PMF	0.33	0.12	0.23	<b>0.98</b>	0.20
Phthalic Acid PMF	0.030	0.014	0.025	0.19	<b>0.96</b>

170

171 Levoglucosan, 4-nitrocatechol, and phthalic acid match well (UC > 0.9). Succinic acid and L-malic acid match less  
172 well, but still have a UC > 0.8. As expected, the UC's for the second standard solution are less good than those for  
173 the first standard solution (which was almost entirely resolved even without PMF).

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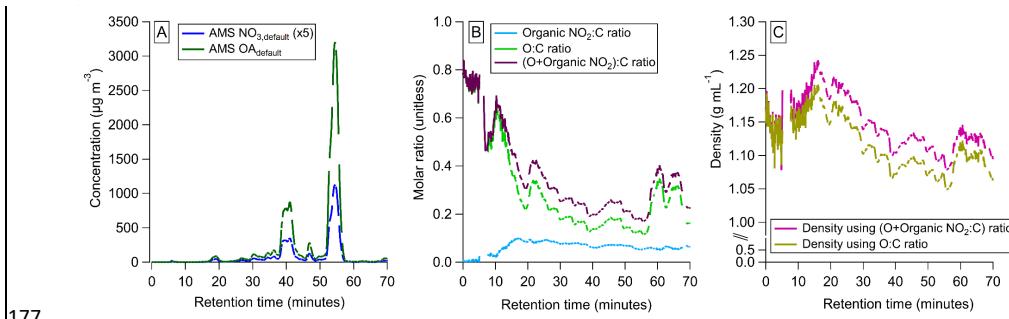
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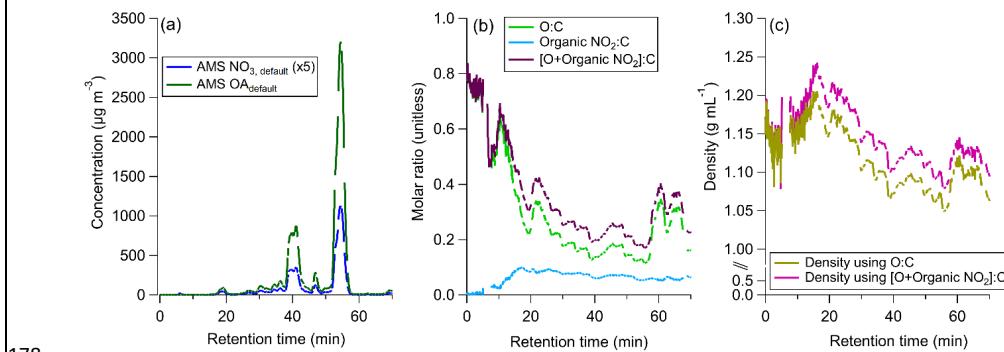
174 **S4  $\beta$ -pinene detailed information: density, molecular identification, PMF solution, and peak fitting**

175 For the SOA samples, the effective density was calculated as described in Sect. 2.5.2, shown in Fig. S8.

176



177



178 **Figure S8. (Aa)** Measured  $\text{NO}_3$  and OA from the AMS when sampling  $\beta$ -pinene +  $\text{NO}_3$  SOA. **(Bb)** Atomic  
179 ratios for organic nitrate : carbon, oxygen to carbon, and oxygen + organonitrates to carbon. **(Cc)** Estimated  
180 density from two approaches.  
181

182 The chromatogram from Claflin and Ziemann (2018) was compared to that measured here (Fig. 7), shown below in  
183 Fig. S9.  
184

185

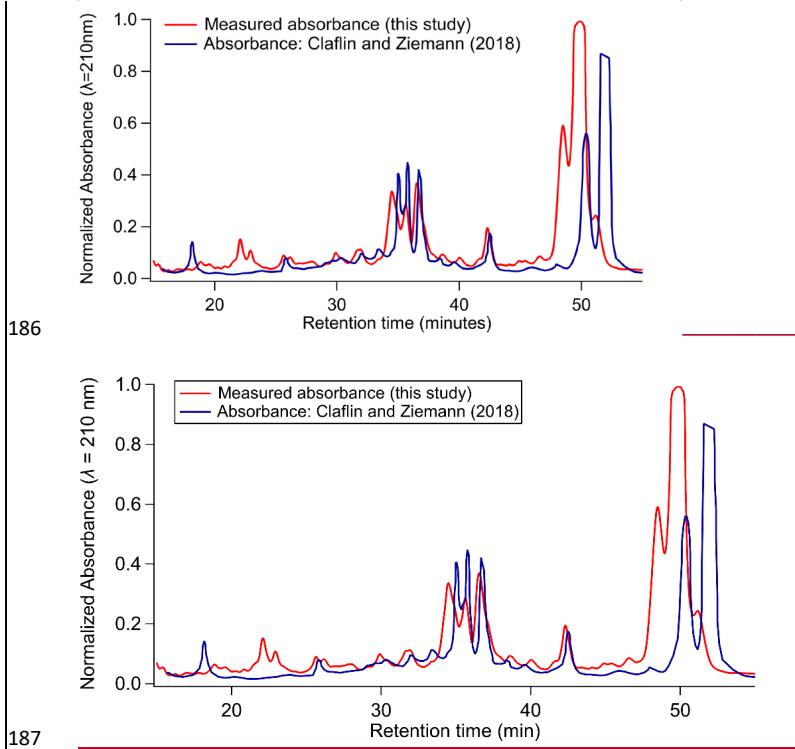
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188 **Figure S9. Comparison to  $\beta$ -pinene +  $\text{NO}_3$  SOA chromatogram measured in Claflin and Ziemann (2018).**

189

190 The chromatograms show the same general shape, although with slightly faster elution for this work. There are some  
 191 notable differences in the results between 20–30 minutes and 45–55 minutes. The final peak in the chromatogram  
 192 from Claflin and Ziemann is the same peak as the largest one measured here (retention time ~50 minutes). This  
 193 suggests that there could be some difference in the HPLC gradient method, or a potential contamination in one of  
 194 the HPLC solvents. Despite that, the overall signals are consistent, and some of the identified species are shown in  
 195 table S6.

196

197 **Table S6. Structure of some known species (from Claflin and Ziemann (2018)), exact (theoretical) mass,**  
 198 **observed mass (measured with EESI+), and mass accuracy (based on EESI instrument multi-ion  $m/z$**

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199 calibration fit).

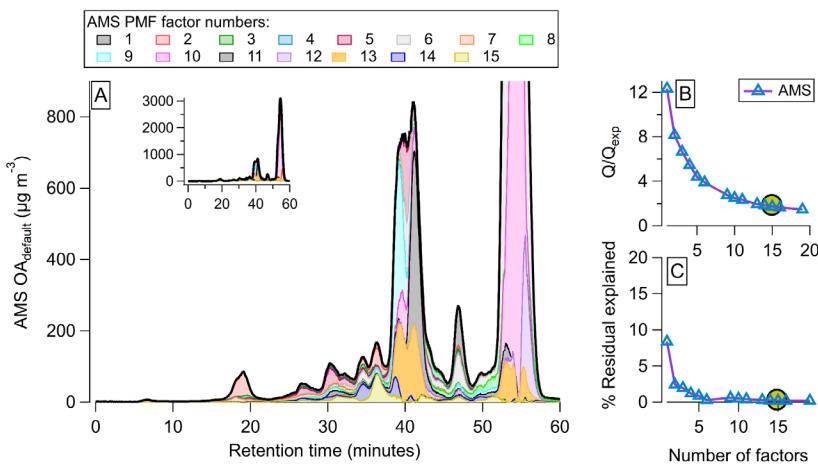
Structure					
MW	245.23	460.48	444.48	460.48	428.48
Exact mass (+Na <sup>+</sup> ) (Da)	268.0797	483.1955	467.2002	483.1955	451.2056
Detected mass (Da)	268.0879	483.1885	467.2032	483.1885	451.2120
Mass Accuracy (ppm)	30.6	-14.5	6.42	-14.5	14.2

200

201

202 PMF was run on the AMS data, shown below for the entire HPLC run (Fig. S10).

203



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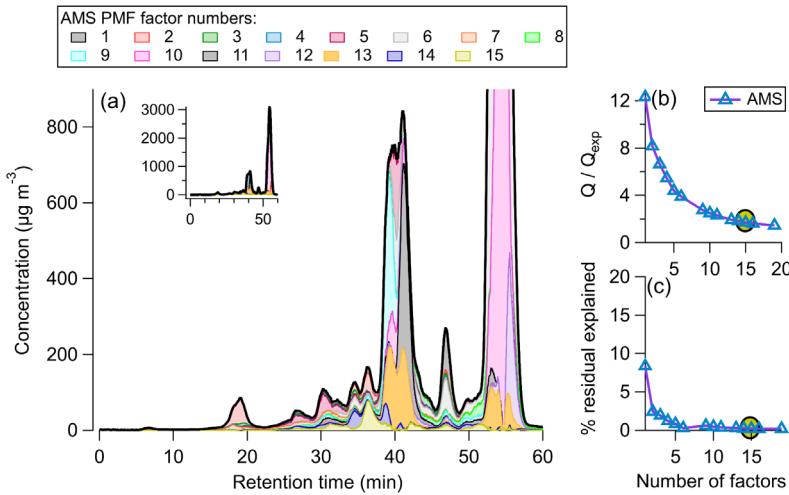


Figure S10. (Aa) stacked plot showing AMS PMF solution time series for the  $\beta$ -pinene +  $\text{NO}_3$  SOA system, with inset showing full scale. (Bb)  $Q / Q_{\text{exp}}$ , with the chosen solution (15 factors) circled. (Cc) Percent of the total sum of the residuals explained, 15 factor solution circled.

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225 collected on the filter. Assuming a 100 % extraction efficiency of SOA, the amount of material injected into the  
226 column can be quantified as such (Eq. S2)

227 
$$\text{Injected mass} = \frac{\text{mass SOA}}{\text{volume ACN}} \times \text{injected volume of solution}$$
 Eq. S2

228 A typical volume of acetonitrile (ACN) used would be ~ 2 mL, therefore the concentration of SOA in ACN would  
229 range from  $313 \mu\text{g mL}^{-1}$  -  $1158 \mu\text{g mL}^{-1}$ . The maximum injected volume is 50  $\mu\text{L}$ , therefore the total injected mass  
230 ranges from 16  $\mu\text{g}$  - 58  $\mu\text{g}$ .

231 To confirm these results, we use the largest peak in the chromatogram ( $m/z$  451.2, retention time ~ 55 min) ←  
232 in an example. According to Claflin and Ziemann (2018), this peak is responsible for ~ 55 % of the total SOA in this  
233 system. Therefore, anywhere from 8.8  $\mu\text{g}$  - 32  $\mu\text{g}$  of the injected mass comes from that peak. However, only 0.55 %  
234 of that mass makes it to the instruments, so the instruments should observe 0.048 - 0.18  $\mu\text{g}$ .

235 The observed AMS mass concentration was roughly  $2000 \mu\text{g m}^{-3}$  using the corrected  $CF_x^A$ . If we assume the peak is  
236 a triangle, we can estimate the area by multiplying the observed peak mass concentration by the total peak elution  
237 time (~ 2 min on average) and dividing by 2. This value is  $2000 \mu\text{g m}^{-3} \times \text{min}$ . The AMS flow was ~  $0.1 \text{ L min}^{-1}$  or  
238  $1 \times 10^{-4} \text{ m}^3 \text{ min}$ , so the AMS sampled ~ 0.2  $\mu\text{g}$ , which is very close to the 0.18  $\mu\text{g}$  estimated above.

239 These injected solution concentrations were able to produce the AMS concentrations observed in Fig. 7,  
240 Fig. S8, Fig. S10, and Fig. S12. For species with a volatility ( $C^*$ ) >  $100 \mu\text{g m}^{-3}$ , there would be substantial  
241 evaporation, > 50 % at equilibrium. While some evaporation would occur for species with a volatility <  $100 \mu\text{g m}^{-3}$ ,  
242 like 4-nitrocatechol in Fig. 4, the SMPS, AMS, and EESI seem to mostly agree.

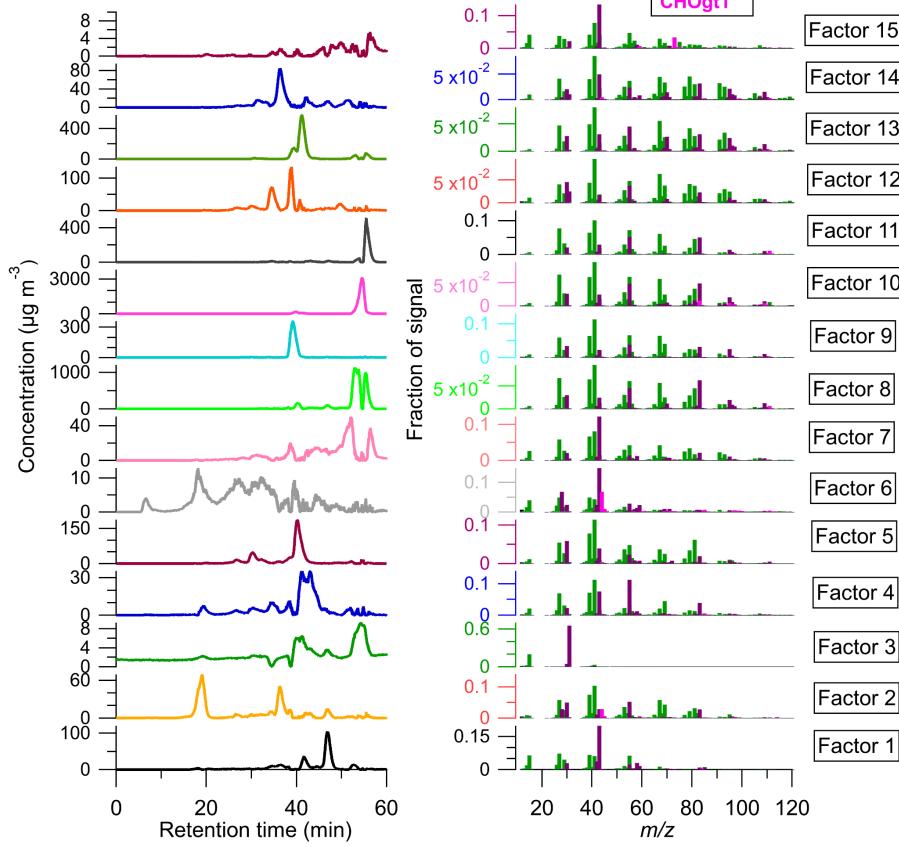
243 It should be noted that, in our setup, < 1 % of the injected mass made it to the mass spectrometers. The use ←  
244 of the collected sample could be optimized further, allowing the analysis of smaller amounts of mass by this method.

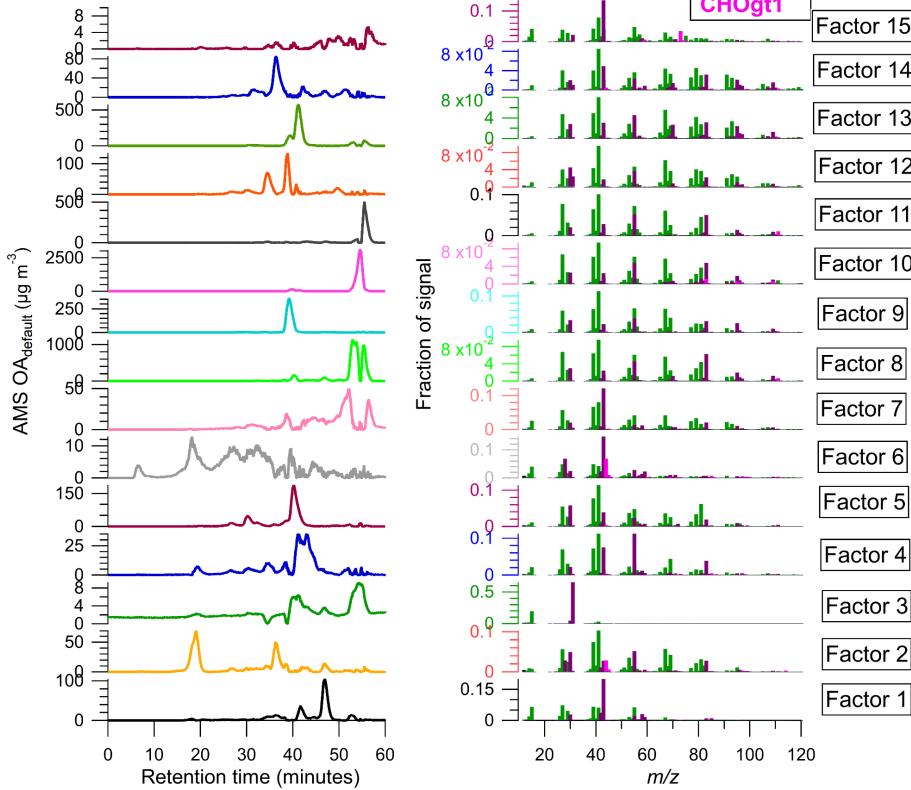
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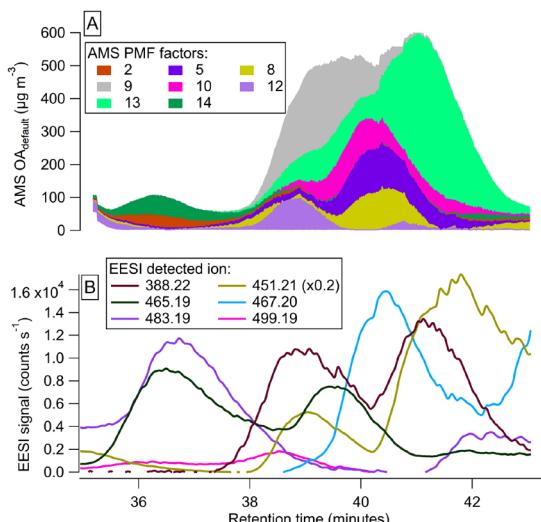
**Figure S11.** (Left) time series of individual PMF factors for the  $\beta$ -pinene +  $\text{NO}_3$  SOA system and (right) HR mass spectra (colored by family) for each factor.

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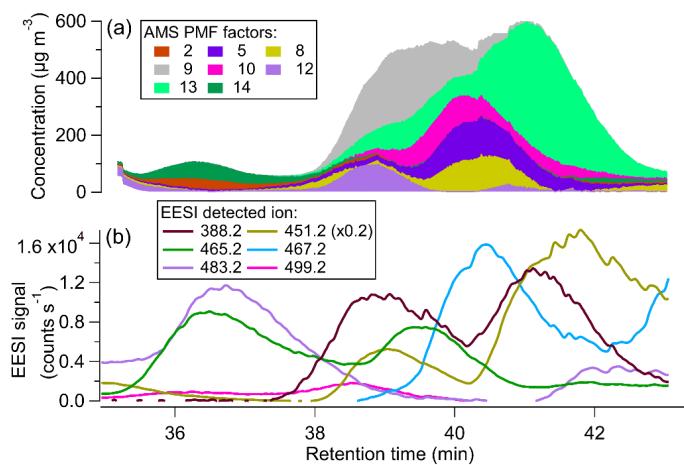
252 Many of the factors have different time series but very similar mass spectra. This suggests that the species fragment  
253 similarly in the AMS (and likely have similar phase states). The SOA products are mostly hydrocarbons with polar  
254 moieties (nitrate, carboxylic acids, ketones, and cyclic ethers). Many of the species retained the nonpolar moiety  
255 from injection to detection (as shown in the CH dominated mass spectra).

256 The peaks eluting from  $\sim 35$  -  $\sim 43$  minutes showed the strongest overlap (and also contained many of the  
257 known  $\beta$ -pinene +  $\text{NO}_3$  SOA products). The time series for this portion of the HPLC run is shown in Fig. S12.

258



259



260

261 **Figure S12. (aA) stacked plot of AMS PMF factors from 35 - 43 minutes and (bB) EESI HR ions time series.**

262

263 As described in Sect. 3.3, EESI HR ions were matched to AMS PMF factors using the shape of the time series' as  
264 well as the retention times. [The EESI HR ions and associated AMS PMF factors are shown in Table S7.](#)

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**Table S7. EESI HR ion and corresponding AMS PMF factor(s)****Formatted:** Font: Bold

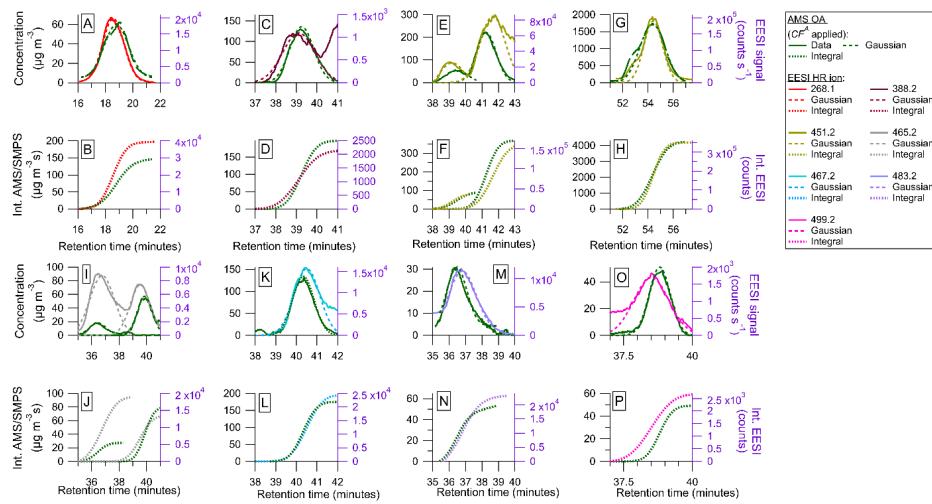
<b>EESI HR ion</b>	<b>Associated AMS PMF factor(s)</b>
<u>268.1</u>	<u>2</u>
<u>388.2</u>	<u>9, 13</u>
<u>451.2 (1)</u>	<u>13</u>
<u>451.2 (2)</u>	<u>13</u>
<u>451.2 (3)</u>	<u>2</u>
<u>465.2 (1)</u>	<u>2</u>
<u>465.2 (2)</u>	<u>10</u>
<u>467.2</u>	<u>5.8</u>
<u>483.2</u>	<u>14</u>

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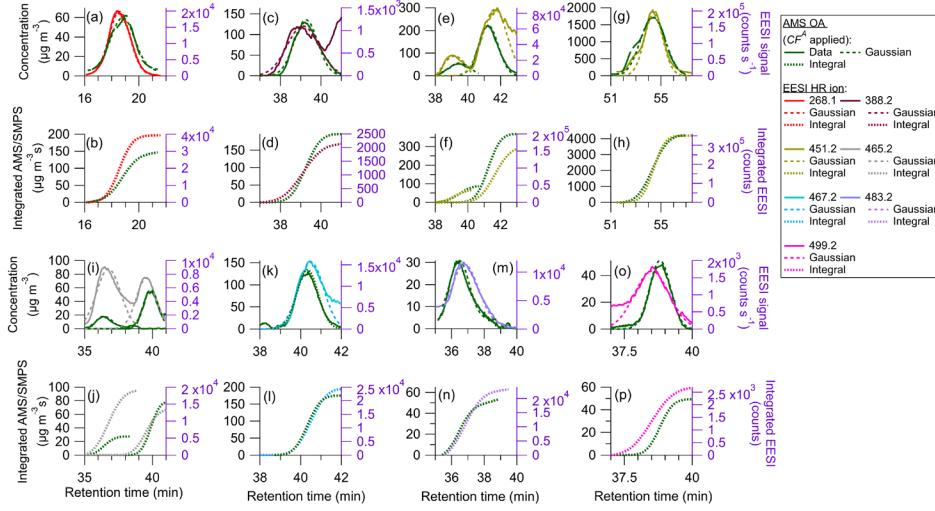
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267 Individual peaks are shown in Fig. S13.

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270

271 **Figure S13.** (a)  $m/z = 268.1$  Gaussians, (b) integrals; (c)  $m/z = 388.2$  Gaussians, (d) integrals; (e) one peak for  
 272  $m/z = 451.2$  Gaussians, (f) integrals; (g) one peak for  $m/z = 451.2$  Gaussians, (h) integrals; (i)  $m/z = 465.2$   
 273 Gaussians, (j) integrals; (k)  $m/z = 467.2$  Gaussians, (l) integrals; (m) one peak for  $m/z = 483.2$  Gaussians, (n)  
 274 integrals; (o)  $m/z = 499.2$  Gaussians, (p) integrals. For the EESI HR ions, the total mass ( $\text{OrgOA} + \text{NO}_3$ ) was  
 275 used in the denominator.

276

277 Not every peak observed in Claflin and Ziemann (2018) was identified here, which is likely due to lack of EESI  
 278 sensitivity to some species and potential decomposition of SOA products (specifically for the trimer identified in  
 279 Claflin and Ziemann (2018)). In contrast, some EESI HR ions that do not correspond to peaks identified in Claflin  
 280 and Ziemann (2018) were detected here, but structures for those species are unknown. All identified individual  
 281 peaks are shown in Fig. S13. As described in Sect. 2.7,  $CF_x^E$  was determined either using the measured SMPS mass  
 282 or the total AMS mass ( $\text{OA}_{\text{Org}} + \text{NO}_3$ ). Fig. S13 shows the AMS OA mass, which was separated by PMF. As  
 283 shown in Fig. S3, the  $\text{NO}_3$  contribution to the total mass was  $\sim 5\%$ . This contribution was added to the denominator  
 284 to calculate  $CF_x^E$  which are reported in table 2 in the main text.

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