

1 **Parameterization of particle formation rates in distinct atmospheric environments**

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17 **Abstract.** Atmospheric particle formation rate (J) is one of the key characteristics in new particle formation (NPF)
18 processes worldwide. It is related to the development of ultrafine particle growth to cloud condensation nuclei
19 (CCN) and, hence, to Earth radiative forcing in global models, which helps us to better understand the impact of
20 NPF on cloud properties and climate change. In this work, we parameterized four semi-empirical J models for 5
21 nm atmospheric particles using field measurements obtained from distinct environments that varied from clean to
22 heavily polluted regions and from tropical to polar regions. The models rely primarily on sulfuric acid as a
23 condensing vapor, condensation sink to account for the vapor loss, and relative humidity for meteorological
24 contribution to J . However, the dependencies between J , condensation sink, and relative humidity are affected by
25 their interlinked relations to sources and sinks of other condensable vapors than sulfuric acid and the potential
26 traffic emissions to the observed size range. The parameterization results showed that our models were able to
27 produce plausible predictions for boreal forest environments, heavily polluted environments, and biogenic
28 environments with high relative humidity. We further tested the models in the global simulation module Tracer
29 Model 5 (TM5, massive parallel version) to simulate particle number size distribution across 14 global atmospheric

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measurement sites. The simulated results showed satisfactory predictions on particle number concentrations for all the tested environments, with significant improvement in the nucleation mode, and better prediction accuracy for Aitken and accumulation modes compared to the binary sulfuric acid-organic vapor model in Riccobono (2014). Our study has successfully provided powerful tools to predict J_5 on a global scale across various environment types using the most essential and more accessible variables involved in the NPF processes. Essentially, this work reinforces the necessity for global research into the investigation of environment-oriented meteorology-involved NPF processes.

1 Introduction

Atmospheric new particle formation (NPF) is a natural phenomenon observed globally (Bousiotis et al., 2021; Brean et al., 2023; Gordon et al., 2017; Kerminen et al., 2018; Nieminen et al., 2018). As particles form and grow on regional scales, they can reach large enough sizes at which they can act as cloud condensation nuclei (CCN) for water vapor to condense onto when forming clouds. This process affects cloud properties (Roldin et al., 2011; Sanchez et al., 2016; Spracklen et al., 2008) and ultimately the global climate depending on the particle numbers, sizes and their chemical compositions (Bellouin et al., 2020; Calvo et al., 2013; Uno et al., 2020). Particle formation rate (J) is an essential parameter describing the NPF intensity, which is often utilized to represent NPF in global models to simulate the effect of NPF on cloud properties and radiative forcing on Earth. To derive a representative parametrization of J for global simulation, we require a broad understanding of NPF in different environments temporally and spatially. That being the case, the atmospheric measurements of both particle number size distributions and NPF precursor vapors are essential to obtain for the atmospheric observations and model developments.

The NPF processes have been investigated and parameterized based on particle formation mechanism theories (Chang et al., 2009; Kulmala et al., 2001; Lehtinen and Kulmala, 2003), field measurements from specific environments, such as pristine boreal forests (Kulmala et al., 2001; Nieminen et al., 2011; Paasonen et al., 2010), urban cities (Salma et al., 2011, 2016, 2021; Salma and Németh, 2019; Zhang et al., 2010), rural area (Lee et al., 2019; Yli-Juuti et al., 2009), marine environment (Zhang et al., 2010), and also chamber experiments (Kirkby et al., 2011; Lehtipalo et al., 2018). In addition to neutral particle formation mechanisms, the ion-induced nucleation is also covered in J parameterizations (e.g. Nieminen et al., 2011; Määttänen et al., 2018). The existing literature primarily focused on the particle activation and survival of nucleation mode particles down to 1.5 nm, involving complex micro-physics of aerosol particles, such as nanometer clusters production, losses due to clusters

coagulation and growth (Bousiotis et al., 2021; Chu et al., 2019; Kerminen et al., 2018; Nieminen et al., 2018). Furthermore, distinct effects on particle formation rates influenced by the same factors were seen under comparable environmental settings. However, these environment-specific models typically have limited application for global simulation implementations that encompass the diverse atmospheric conditions on Earth.

To simulate particle formation rates, one usually starts from the nucleation mode size range. Global modelers have been facing great challenges in simulating nucleation mode particles because large-scale models have limited capabilities in treating the complicated aerosol dynamics taking place in the sub-5 nm particle size range. The formation rate at 5 nm is shown to be important because after sizes of about a few nm in diameter, particle growth rates show relatively limited variability in different environments (Kulmala et al., 2022a, b, 2023). In addition, we currently do not have a good enough theoretical understanding on the processes dictating particle growth rates at the smallest sizes, nor the survival of such particles from coagulation scavenging (e.g. Cai et al., 2022; Tuovinen et al., 2022; Marten et al., 2022).

The essential parameters for J parameterizations should include at least one type of precursor vapor, some may also cover meteorological parameters, and the sinks for vapors and particles. For instance, sulfuric acid (H_2SO_4) as the most known precursor vapor plays a critical role in particle formation and growth processes due to its low volatility (Kulmala et al., 2004; Myllys et al., 2019). In the earlier parameterizations of NPF mechanism, J correlated (linearly or squared) with H_2SO_4 concentrations in various environments (Paasonen et al., 2010). In terms of meteorology, air temperature (T), relative humidity (RH), global solar radiation (GRad), wind speed (WS), and wind direction influence the particle formation rates in certain environments as well (Laarne et al., 2022; Salma et al., 2021; Zaidan et al., 2018). The variation of T can influence the precursor vapor formation and stability of NPF processes: a higher T can enhance the biogenic emissions that participate in particle formation in a boreal forest (Dada et al., 2017; Nieminen et al., 2015), while a lower T favored the H_2SO_4 -amine clusters stability in a megacity (Deng et al., 2020). RH can impact the precursor vapor formations as well as the aerosol formation rates (Ding et al., 2021; Hellmuth, 2006). The variation of RH is dependent upon T so that the rise of T during daytime increases the planetary boundary layer height (PBLH), which in turn dilutes the air mixture and decreases the RH (Liu et al., 2018) as well as particle number concentrations (Mazon et al., 2016) in the atmosphere. For condensable vapor loss, we usually include the term condensation sink (CS), which describes the loss rate of condensable vapors to aerosol particles, and it typically declines before an NPF event starts. H_2SO_4 concentrations, on the other hand, increase due to the reduction in CS, which means that the condensable vapors are not lost onto the aerosol particles as efficiently as they would be at greater CS values (Hellmuth, 2006; Kulmala et al., 2012).

89 In general, the developed J models underestimate the observed particle number concentration, which may be
 90 attributed to NPF schemes being poorly represented in these models. Many J parameterization works were
 91 conducted focusing on the formation mechanisms from sulfuric acid (Paasonen et al., 2010), sulfuric acid-water
 92 (Määttänen et al., 2018), sulfuric acid-ammonia (Glasoe et al., 2015), and sulfuric acid-organic vapor (Paasonen et
 93 al., 2010; Riccobono et al., 2014). However, some models are likely applicable only to certain types of environment,
 94 or they primarily cover the microphysics of the particle nucleation at sub-3 nm range, where the nucleated clusters
 95 face higher instability due to the higher evaporation rates than condensation rates (Deng et al., 2021; Wang et al.,
 96 2011). Bergman (2022) attempted an organic-vapor-based NPF scheme in addition to the commonly used binary
 97 water-sulfuric-acid-based scheme to simulate global particle formation and number concentrations. This scheme
 98 improved the simulated number concentrations across the observation stations, although they were still
 99 underestimated compared to the observations, suggesting that the parameterization of early growth of particles to
 100 5 nm diameter still requires improvement.

101 To predict the particle formation rate at 5 nm originating from NPF and subsequent growth, as well as to understand
 102 and predict the climatic impacts caused by NPF and initial growth in global scale, we parameterize particle
 103 formation rates (J) at 5 nm using combined measurement data from six different environments: Hyytiälä (boreal
 104 forest close to rural environment, Finland), Beijing (megacity, China), Värriö (remote boreal forest, Finland),
 105 Budapest (urban, Hungary), Agia Marina Xyliatos (rural, Cyprus) and Manacapuru (Amazonian basin, Brazil). The
 106 parameterizations of J were based on the analysis of atmospheric particle number-size distributions. Sulfuric acid
 107 concentrations, RH and CS are the main input variables in the parameterization models. By including information
 108 from various types of environments, we will be able to demonstrate whether our models can adequately explain the
 109 formation rate of 5 nm particles on a wider environmental scale. The parameterized models are then incorporated
 110 into EC-Earth models to simulate particle formation rates in the global scale (the European community Earth-
 111 System Model, EC-Earth, chemistry transport model TM5: Tracer Model 5, version TM5-chem-v3.0, details in
 112 supplement) (Huijnen et al., 2010).

113 This work aims to provide an effective tool for global particle formation rate estimations. Our parameterizations
 114 have three main features: (1) the number of inputs is limited to be the most essential parameters involved in NPF
 115 process, (2) they do not involve complex microphysics at particles smaller than 5 nm, and (3) they cover a wide
 116 range of environment types. These features will enhance the applicability of the parameterizations for the purpose
 117 of global model application.

118 **2 Measurement locations and instrumentation**

119 This study includes measurements from six different sites representing different environmental conditions. A
120 summary for all locations and the instrumentation used is given in Table S1. Figure 1 shows the map of the
121 measurement sites included in this study.

122 **2.1 Measurement sites**

123 **2.1.1 Rural boreal forest environment: Hyytiälä, Finland**

124 The measurement data were obtained from the SMEAR II-station (Station for Measuring Ecosystem–Atmosphere
125 Relations), situated in a Scots pine (*Pinus sylvestris*) forest in Hyytiälä (61.1 N, 24.17 E; 181 m (a.s.l.); Hari and
126 Kulmala, 2005), southern Finland. This measurement site is described as having a rural regional background with
127 minimal anthropogenic emission. Hyytiälä data covered the period from 21 March 2016 to 18 August 2019.

128 **2.1.2 Remote sub-arctic boreal forest: Värriö, Finland**

129 The SMEAR I measurement station (67°45′31″ N, 29°36′41″ E, 390 m a.s.l.) was built on the top of Kotovaara hill
130 located in north-eastern Finland. Similar to Hyytiälä, the site is also a rural background covered mainly by Scots
131 pine (*Pinus sylvestris*) forest located at the north side of the Värriö fell range. However, it is affected by potential
132 polluted air mass that comes from Kola peninsula rather than local industrial pollutants. A detailed description of
133 SMEAR I station can be found in (Kyrö et al., 2014). The data used from Värriö were from 5 April to 13 August
134 2019.

135 **2.1.3 Polluted megacity: Beijing, China**

136 In Beijing, the measurements were performed at the west campus of the Beijing University of Chemical Technology
137 (BUCT, 39.94° N, 116.30° E, 20 m a.s.l.). The sampling took place from outside the window on the 5th floor of the
138 university building close to a street with busy traffic. For more details on the description of BUCT measurement
139 site, see Liu et al., 2020. The data were available from 29 May 2018 to 3 April 2019.

140 **2.1.4 Urban site: Budapest, Hungary**

141 The measurements took place at the Budapest platform for Aerosol Research and Training (BpART) Laboratory
142 (47.47° N, 19.06° E; 115 m a.s.l.) of the Eötvös Loránd University situated on the bank of the river Danube. The
143 site represents a mixed average atmosphere of the city center (Salma et al., 2016). The data were obtained from 22
144 March to 17 April 2018.

145 **2.1.5 Mediterranean rural site: Agia Marina, Cyprus**

146 The measurements were conducted at the Agia Marina Xyliatou (AMX) station (35.03° N, 33.05° E; 532 m a.s.l.)
147 of the Cyprus Atmospheric Observatory (CAO). The site represents a rural background location situated at the
148 foothills of Troodos mountains, with agriculture land in the vicinity. The data were obtained between 22 February
149 and 3 March 2018. For more details about the site, see e.g. Baalbaki et al., (2021).

150 **2.1.6 Amazonian basin: Manacapuru, Brazil**

151 The Manacapuru measurement site was in a pastureland 70 km west of Manaus, Brazil, in central Amazonia. This
152 site receives air mass from various resources, including rural, biogenic and anthropogenic from the nearby
153 municipality (Manaus). The trace gases and meteorological measurements were performed during the
154 GoAmazon2014/5 campaign at the T3 site (3.2133° S, 60.5987° W 50 m a.s.l.), 10 km northeast of Manacapuru,
155 Brazil (Martin et al., 2016; Schiro et al., 2018). A more detailed description of the measurement site can be found
156 in (Myers et al., 2022). The data covered the time period from 22 August 2014 to 9 October 2014.

157 **2.2 Instrumentation**

158 **2.2.1 Sulfuric acid measurements and proxies**

159 H₂SO₄ concentrations were measured at all sites, except for the Amazonian basin, using a Chemical Ionization
160 Atmospheric Pressure interface Time-of-Flight spectrometer (CI-API-ToF) (Eisele and Tanner, 1993; Jokinen et
161 al., 2012) with NO₃⁻ as the reagent ion and analyzed using tofTools package based on MATLAB software
162 (Junninen et al., 2010). In the Amazonian basin, H₂SO₄ concentrations were measured using a selected ion chemical
163 ionization mass spectrometer (SICIMS), see Myers et al., (2022) for more details. The H₂SO₄ concentration
164 measurements were taken from different levels ranging from ground level up to 35 meters above ground level. The
165 CI-API-ToFs were calibrated uniformly before the measurement in each location following the technique described

by (Kürten et al., 2012), except for the Amazonian basin where the selected ion chemical ionization mass spectrometer (SICIMS) was calibrated following the scheme described in Mauldin III et al. (1998). To increase the applicability of our derived parameterization, H₂SO₄ proxy data from Hyytiälä and Beijing were included as an additional testing data set. The proxy data were calculated using the proxy specific for the boreal forest environment and polluted megacity developed by Dada et al., (2020). For Hyytiälä, the sulfuric acid proxy data ranged from 22 August to 25 December 2016, and 8 March 2018 to 26 February 2019, denoted as Hyytiälä_{SAprx}; For Beijing, the time period was from 15 March to 3 April 2019, denoted as Beijing_{SAprx}. The subscript “SAprx” (SA as in sulfuric acid) in Hyytiälä_{SAprx} and Beijing_{SAprx} indicates that the datasets utilize H₂SO₄ concentration from proxies as input for the testing dataset.

2.2.2 Particle number size distribution

The particle number size distribution (PNSD) measurements were obtained from different types of setups in each site. Hyytiälä: twin-Differential Mobility Particle Sizers (DMPS; Aalto et al., 2001); Värriö: Differential Mobility Particle Sizers (DMPS) (Jokinen et al., 2022); Beijing: Particle Size Distribution (PSD) system with a nano-Differential Mobility Analyzer (DMA) and an Aerodynamic Particle Sizer (APS) (Zhou et al., 2021); Budapest: flow-switching-type DMPS (6 – 1000 nm; Salma et al., 2016); Cyprus: Neutral cluster and Air Ion Spectrometer (NAIS) and Scanning Mobility Particle Sizer (SMPS; Baalbaki et al., 2021); Amazonian basin: the measurements were conducted using SMPS (10 – 1000 nm). It is important to note that we do not aim to compare the PNSD measurements from all the chosen sites. Instead, the PNSD measurements were used to calculate the formation rates based on changes in particle number concentrations under local conditions.

2.2.3 Meteorological variable

The meteorological variables included in this study are relative humidity (RH, %) and ambient temperature (T , °C). In Hyytiälä, RH and T were measured at 16.8 m using Rotronic MP102H RH sensor (Rotronic Hygromet MP102H with Hygroclip HC2–S3, Rotronic AG, Bassersdorf, Switzerland); In Värriö, RH and T were measured by a Rotronic MP106A captive sensor; In Beijing, RH and T were monitored by Vaisala weather station (AWS310); In Budapest, RH and T were monitored using Vaisala HMP45D temperature and humidity probe, and Vaisala WAV15A anemometer located on on-site of the BpART Lab; In Cyprus, RH and T were measured by a meteorological station in a nearby village (35.01° N, 33.05° E), 2.85 km away from the measurement site; In Amazonian basin, RH and T were measured at the Atmospheric Radiation Measurement (ARM) user facility.

194 2.3 Data analysis

195 2.3.1 Calculation of particle formation rates

196 To develop more inclusive and generalized models, the parameterization included data from both NPF event days
197 and non-event days. This approach recognizes that the production of atmospheric secondary particles from non-
198 NPF events (days with no apparent particle growth) is becoming more significant in a world with growing
199 anthropogenic influence (Kulmala et al., 2022a). Such a measure would increase the applicability of our models on
200 a global scale.

201 The observed particle formation rates (J_5) at 5 nm were calculated from the measured PNSD according to Equation
202 1 (Kulmala et al., 2012).

$$203 J_{dp} = \frac{dN_{dp}}{dt} + CoagS_{dp} * N_{dp} + \frac{GR}{\Delta d_p} * N_{dp} \quad (\text{Eq. 1})$$

204 The first term dN_{dp}/dt is the change in concentration in the size bin, 5–9 nm. Ideally, this term, as well as the
205 concentration of particles within the size range, N_{dp} , in the following terms, are associated with the growth of
206 particles formed by atmospheric NPF past 5 nm; however, especially in traffic-related environments, they may also
207 have an unknown contribution by direct particle emissions to this respective size range (Okuljar et al., 2021;
208 Rönkkö et al., 2017). The second term $CoagS_{dp}$ is the coagulation sink, which describes the 5-9 nm particle losses
209 due to coagulation with larger particles calculated from the PNSD at each measurement site (Kulmala et al., 2012).
210 The third term describes the loss of particles due to their growth out of the size bin. Here, we calculated the growth
211 rates (GR) of 5–9 nm particles using the maximum concentration method (Kulmala et al., 2012) for days classified
212 as NPF event days as described by Dal Maso et al., (2005). The GR for non-event days was approximated using
213 the normalized PNSD from the sum of non-NPF events at each site. Such approximation is validated for several
214 locations as a ‘quiet NPF’ occurs with the similar GR as that on NPF event days (Kulmala et al., 2022a).

215 2.3.2 Extrapolation of particle formation rates

216 For Budapest and the Manacapuru (Amazonian basin), the particle formation rates were calculated from PNSD
217 measurements at 6 nm and 10 nm, respectively. Therefore, we obtained J_5 by extrapolating from J_6 and J_{10}
218 respectively. The J_5 extrapolation followed the analytical formula derived by Kerminen and Kulmala (2002). We
219 extrapolated J_5 from J_6 for Budapest. For Manacapuru, the extrapolations were done separately for J_{10} (wet season)
220 and J_{14} (dry season), due to the particle size limit of the measurement instrument.

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221 **2.3.3 Condensation and coagulation sink (CS and CoagS)**

222 The CS and CoagS were calculated from the measured PNSD data for each site using the method proposed by
223 Kulmala et al., (2012). To ensure the comparability between all locations, both CS and CoagS were calculated
224 without the correction for hygroscopic growth. There are several ways to determine the hygroscopic growth factors
225 in CS and CoagS calculations. Laakso et al., (2004) developed parameterizations for Hyytiälä solely based on the
226 meteorological conditions and the aerosol composition in Hyytiälä, which results in the inapplicability of that
227 method to other sites. In the Supplementary of Baalbaki et al., (2021), Figure S4 shows the CS with hygroscopic
228 correction is about 1.1 – 1.3 times higher than dry CS, which would result in an overestimation on CS for the case
229 in Cyprus. Petters and Kreidenweis (2007) introduced the single hygroscopicity parameter κ (kappa), which can be
230 derived from Humidified Tandem Differential Mobility Analyzer (HTDMA) or cloud condensation nuclei counter
231 measurements or based on aerosol chemical composition obtained from instruments such as the Aerosol Chemical
232 Speciation Monitor (ACSM) or Aerosol Mass Spectrometers (AMS). In other locations, since organics are typically
233 the dominant component of aerosol mass in continental areas or marine polluted areas (Chen et al., 2022) and are
234 less hygroscopic than inorganics, one can expect an underestimation of CS similar to the one reported in (Baalbaki
235 et al., 2021). As a result, we omitted the hygroscopic growth impact for the chosen measurement sites to harmonize
236 the data composition and the later model analysis.

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Part 1/2.

237 **2.3.4 Datasets**

238 The parameterizations were developed using the combined dataset from all six measurement sites in hourly time
239 resolution. Data points were selected considering detection limit of the instruments and therefore, the filters were
240 set to be $J_5 > 1 \times 10^{-5} \text{ cm}^{-3} \text{ s}^{-1}$, H_2SO_4 concentration $> 5 \times 10^3 \text{ cm}^{-3}$, $\text{RH} \in [0, 100] \%$ and $\text{CS} > 1 \times 10^{-5} \text{ s}^{-1}$. The complete
241 dataset was afterwards randomly resampled into a training set (75% from the complete dataset) and a testing set
242 (25% the rest of the complete dataset) for parameterization. In model testing, we included two additional inputs
243 from H_2SO_4 concentration proxies developed by Dada et al., (2020) from Hyytiälä and Beijing. The detailed number
244 of data points per site are shown in Table 1. The data distribution and comparison of each input variable are
245 displayed in Figure S2, where the overall variations of the input variables across the six sites are distinct in their
246 range and intensity, which pronounces the inclusivity of model training for a wider application in global
247 environments.

248

249 **3 Parameterization of J_5**

250 **3.1 Derivation of parameterization models**

251 We derived the parametrized J_5 based on the input variables (H_2SO_4 , RH, CS), which were chosen based on field
252 observations that highlighted their roles in the particle formation mechanism across various environments (Baalbaki
253 et al., 2021; Dada et al., 2020; Kerminen et al., 2018; Myers et al., 2022; Salma et al., 2016, 2021; Yan et al., 2021).
254 It has been discovered that NPF events occur favorably under lower RH, for example in boreal forests (Dada et al.,
255 2018; Yao et al., 2018), Mediterranean regions (Debevec et al., 2018), from CLOUD chamber experiment
256 (Duplissy et al., 2016) and model studies (Hamed et al., 2011). RH was shown to be seasonally related to cloudiness
257 and global radiation, so that a decreasing global radiation can lead to an increased RH and cloudiness within the
258 troposphere (Ruosteenoja and Räisänen, 2013). To reduce the model complexity, we opted to use RH as an indirect
259 indicator of global radiation. A lower CS facilitates the occurrence of NPF events even in contrasting environments
260 with distinct types of condensable vapor. For example, CS is a measure of a sink for anthropogenic vapors in a
261 megacity (Wang et al., 2011) and for biogenic vapors in a clean boreal forest (Dada et al., 2017; Tuovinen et al.,
262 2020), as well as a sink for growing sub-5 nm clusters and particles (Kulmala et al., 2017). When combined with
263 H_2SO_4 as an input variable, the evidently important sink effect of a pre-existing particle population on the ambient
264 H_2SO_4 concentration is implicitly transferred from CS to H_2SO_4 in our parameterization. Indirectly, CS may also
265 be associated, either causally or not, with 1) emissions or sinks of vapors other than H_2SO_4 participation in NPF or
266 particle growth, and 2) primary particle emissions from traffic, which would influence particle formation rates
267 estimated from observations using eq. 1. Furthermore, since we omit the influence of hygroscopic growth of
268 particles on CS, a fraction of real sink effect of CS is implicitly transformed to the variable RH in our
269 parameterization.

270 We tested with T as an input variable during model derivation and training. However, the modelled results did not
271 show improvement compared to the current parameterization, suggesting that T provided redundant information
272 for describing particle formation in the context of our model's global application.

273 Other than sulfuric acid, highly oxygenated organic molecules (HOMs) and ammonia (NH_3) have been discovered
274 to play a significant role in particle formation process (Bianchi et al., 2019; Lehtipalo et al., 2018). The possible
275 cluster types may include $\text{H}_2\text{SO}_4\text{-NH}_3\text{-H}_2\text{O}$ (Yu et al., 2018) and HNO_3 -related clusters such as $\text{HNO}_3\text{-H}_2\text{SO}_4\text{-NH}_3$
276 in the upper tropospheric particle nucleation (Wang et al., 2020, 2022, 2023). Yet, we are unable to include HOM
277 nor NH_3 concentrations owing to limited data availability from the chosen measurement sites. So far, long-term

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Part 2/2.
Also for RC2, Comment 2.

Commented [XL5]: Modification for RC2, Comment 6.

measurements (> 1 year) of HOMs, matching the time range covered by other variables, are only available in Hyytiälä from a CI-API-ToF mass spectrometer. However, this is not the case at other sites, limiting our ability to have simultaneous HOMs data across all environments included in this study. Similarly, the NH₃ concentrations either did not cover the same time period as other variables or were unavailable for the other environments.

3.1.1 Different versions of the parameterization models

The derived model functional forms are as follows:

Model 1 (the baseline model, Eq. 2) presents the simplest particle formation mechanism based solely on the abundance of the precursor vapor H₂SO₄ concentrations in the atmosphere. The coefficient k_1 serves as a scaling coefficient that represents the activation rate of clusters in the presence of H₂SO₄ molecules during cluster formation (Kulmala et al., 2006; Paasonen et al., 2010).

$$J_5 = k_1 \times [\text{H}_2\text{SO}_4] \quad (\text{Eq. 2})$$

Model 2 (Eq. 3) introduces RH in addition to model 1 to partially represent the effect of the changing meteorological conditions relating to the global radiation and ambient water vapor content on J_5 in general in different types of environments (Dada et al., 2017; Hamed et al., 2011; Li et al., 2019). The coefficient k_2 serves as a scaling coefficient and shown as the activation efficiency of the nucleated clusters.

$$J_5 = k_2 \times [\text{H}_2\text{SO}_4] \times \text{RH}^{k_{\text{RH}}} \quad (\text{Eq. 3})$$

Model 3 (Eq. 4) includes, in addition to model 2, the factor CS. As discussed above, in our parameterization CS is connected not only to the sink of newly formed particles prior to their growth past 5 nm, but possibly also to sinks or sources of vapors other than H₂SO₄ participating in particle formation and growth and, in polluted environments, to sub-10 nm particle emissions from traffic. The coefficient k_3 serves as a scaling coefficient for the activation and survival efficiency of the nucleated clusters.

$$J_5 = k_3 \times [\text{H}_2\text{SO}_4] \times \text{RH}^{k_{\text{RH}}} \times \text{CS}^{k_{\text{CS}}} \quad (\text{Eq. 4})$$

Model 4 (Eq. 5), additionally accounts for the formation of H₂SO₄ multimers in the gas phase prior to cluster formation as assumed by the kinetic theory (McMurry and Friedlander, 1979), the coefficient k_{SA} represents the number of H₂SO₄ molecules (2, 3, 4, etc...). Therefore, k_4 in this case is not the activation coefficient anymore but includes both the collision frequency and the probability of a stable particle formation after the collision (Sihto et al., 2006; Weber et al., 1996).

$$J_5 = k_4 \times [\text{H}_2\text{SO}_4]^{k_{\text{SA}}} \times \text{RH}^{k_{\text{RH}}} \times \text{CS}^{k_{\text{CS}}} \quad (\text{Eq. 5})$$

Commented [XL6]: Additional text considering the overall Referee Comments.

3.2 Model training results

To derive a parametrized J_5 based on precursor and other input variables from the training dataset, we used the “*fmincon*” optimization algorithm in MATLAB to retrieve the values of each coefficient (k_1 - k_4 , k_{SA} , k_{RH} and k_{CS}) from the training dataset. The coefficients obtained for each of the models can be found in Table 2. The derived models with the optimized coefficients were applied to the testing datasets and compared with the observed J_5 and the parametrized J_5 . We evaluated the performance of each model based on the data distribution, the resulting deviation from observation and its uncertainty. To maintain the global model’s simplicity, the parameterization covered both daytime and night-time data for all sites in all models.

Figure S3 presents the measured to modelled J_5 from model 1-4 using training dataset from six measurement sites, including the slopes and coefficient of determination (R^2). Overall, by comparing model 1 (Fig. S3a) and model 2 (Fig. S3b), we observed an improvement in the model performance with the inclusion of RH. The R^2 value improved from 0.28 to 0.44, and the slope increased from 0.29 to 0.56. This observation confirmed the importance of considering meteorological impact when parameterizing J_5 . By further including CS in model 3, the model improved further (Fig. S3c), with the R^2 increasing from 0.44 to 0.49, and the slope from 0.56 to 0.62. To further introduce the kinetic theory and the formation of H_2SO_4 dimers and other multimers, we added an exponent over H_2SO_4 in model 4 (Fig. S3d). This addition showed a further improved correlation and slope between the measured or modeled data for the training datasets ($R^2 = 0.57$, slope = 0.76). In subsequent testing, model 4 generally outperformed the other models (see section 4, Fig. 2).

3.3 Model evaluations

3.3.1 MAE and RMSE

We computed the mean absolute errors (MAE), root mean square errors (RMSE) for each model using the testing dataset to gain a better understanding of the models’ performance. The numerical values of MAE and RMSE are given in Table S3.

The MAE calculation equation is as follows:

$$MAE = \frac{1}{n} \sum_{i=1}^n |y_i - \hat{y}_i| \quad (\text{Eq. 6})$$

where n is the number of data points (here it is the total number of data points from testing set, see Table 1), y_i is the observed value, and \hat{y}_i denotes the predicted value. MAE measures the accuracy of models' prediction power, by quantifying the average magnitude of errors between observed and predicted values (Chai and Draxler, 2014). A lower model error is manifested by a lower MAE value. The RMSE is calculated as the square root of the difference between the measured (y_i) and predicted (\hat{y}_i) J_5 values normalized by the number of data points.

$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (y_i - \hat{y}_i)^2} \quad (\text{Eq. 7})$$

RMSE also measures the average magnitude of errors of models. However unlike MAE, RMSE squares the errors giving greater weight to larger errors and penalizing them more heavily (Chai and Draxler, 2014). Therefore, RMSE values reveal whether the models' performances are highly influenced by large prediction errors. Similar to MAE, lower RMSE values indicate better model performance. Figure S4 (upper panel) depicts a declining trend of the overall MAE from models 1 to 4 (Eqs. 2-5). For the environmental types investigated in this study, the MAE values of the four models from all sites are lower than 1, indicating that the mean differences on the magnitude for J_5 are minor when utilizing the parameter settings from our models. However, Budapest stands out due to the apparent higher MAE, potentially highlighting the distinct NPF mechanism in Budapest compared to the other sites as well as the seasonal limitations on its data (spring 2018 only). The RMSE values increased as more parameters were added into the model, peaking for model 3 (Fig. S4, lower panel), even though model 3 can predict J_5 for multiple types of environments on a satisfactory level. We can see that from model 1 to model 2, the inclusion of RH increased the model errors more compared to the addition of CS from model 2 to model 3. However, the RMSE values dropped significantly when H_2SO_4 was allowed to vary with an exponent k_{SA} in model 4 in the presence of both RH and CS. Based on the results summarized above, models 3 and 4 (Eqs. 4 & 5) seem to be the most promising for global J_5 prediction among all model types owing to their low MAE values. However, the lower RMSE for model 4 showed its outperformance to model 3.

356 **3.3.2 Akaike Information Criterion**

357 The Akaike Information Criterion (AIC) is a statistical measure that helps to evaluate the goodness-of-fit of a
358 statistical model. We use AIC as an evaluation tool because it can evaluate models with different number of
359 parameters and complexities, ensuring a balanced assessment. Eventually, it allows us to select the model with the
360 best balance between the model complexity and goodness-of-fit. The parameters used to calculate the AIC for each
361 site are shown in Tables S4-S6. A lower AIC score indicates a superior goodness-of-fit and a lower tendency for
362 model overfitting. The relative likelihood term (L_i , $L_i = e^{(AIC_{min} - AIC_i)/2}$), calculated from AIC scores, reflects the
363 likelihood that the i th model minimizes information loss as compared to the model with the lowest AIC. A relative
364 likelihood of 1 suggests that the model outstands other models in minimizing information loss. For boreal forest
365 environments (Table. S4), and urban environments (Table. S5), both models 1 and 4 minimized information loss
366 the most. For rural regions, model 4 (Eq. 5) performs the best (Table. S6). Compared to the baseline model (model
367 1, Eq. 2), we find that H_2SO_4 is a more powerful parameter than RH or CS in all environments. However, in
368 Manacapuru, including RH and CS shows clearly an improved predictive accuracy in model 4 (Eq. 5).

369 **4 Results and discussion**

370 **4.1 Parameterization testing results**

371 The scatterplots (Fig. 2) demonstrate the overall performance of the parameterizations from the 4 models (Eqs. 2-
372 5) using the testing dataset. The overall and site-specific Pearson's coefficients, slopes from robust linear fit
373 between the measured and modelled J_5 , as well as the number of data points from the testing dataset, can be found
374 in Table 3. Overall, r increased significantly for model 2 and 3 (Fig. 2b & 2c, $r = 0.69$, $r = 0.71$) compared to model
375 1 (Fig. 2a, $r = 0.55$) as we include relative humidity and condensation sink as model parameters. Model 4 provides
376 the best linear fit results, implying that the model can predict an overall reliable estimation on J_5 in all the
377 investigated environment types (Fig. 2d, $r = 0.78$). It is notable that with the combined data sets, the condensation
378 sink receives a positive exponent in models 3 and 4 ($k_{CS} = 0.56$ and 0.67 , respectively), likely due to its association
379 with concentrations of other condensable vapors than H_2SO_4 and traffic emissions.

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380 4.1.1 Boreal forests: Hyytiälä and Värriö

381 Given the boreal forest background, Hyytiälä and Värriö exhibited comparable variations in the distribution of
382 modelled J_5 values from the four model types. As shown in Figure 3, model 3 ((a3), (b3), (c3)) and model 4 ((a4),
383 (b4), (c4)) illustrated a more centered data distribution between the modelled and measured J_5 , which demonstrates
384 a potential favoring of NPF under low RH conditions, likely associated with increased global radiation for the
385 boreal forest environment (Dada et al., 2018; Hamed et al., 2011) and lower sinks for vapors and growing sub-5
386 nm particles (see Section 3.1). Notice that the mean H_2SO_4 concentration in Värriö is about twice as high as that in
387 Hyytiälä, opposite to CS which is clearly lower in Värriö (Table S2). The low CS in Värriö compared with Hyytiälä
388 is primarily due to the lower emission rate of the regional precursor vapors (e.g. Tunved et al., 2006), leading also
389 to the lower observed NPF event frequencies (Kyrö et al., 2014; Neefjes et al., 2022). We must note that the
390 Hyytiälä data spanned three years, containing more data points for model training, whereas the Värriö data only
391 covered the period from April to August 2019, excluding the entire cold season when H_2SO_4 concentrations are
392 significantly lower than those during the warm season (Jokinen et al., 2022). As a result, our model 3 (Fig. 3(a3),
393 (b3), (c3)) and 4 (Fig. 3(a4), (b4), (c4)) can predict J_5 for boreal forest environment on a satisfactory level, including
394 the possibility to use the estimated H_2SO_4 concentration from proxies as input. Nevertheless, limitations regarding
395 precursor vapor production rate could potentially influence the prediction accuracy.

396 4.1.2 Urban-influenced: Beijing and Budapest

397 In anthropogenic emissions dominated region, such as Beijing, the measured and modelled J_5 are well aligned
398 around the 1:1 line using model 3 (Fig. 3(d3), (e3), (f3)), and model 4 (Fig. 3 (d4), (e4), (f4)). In Beijing, a polluted
399 megacity, the dominating precursor type has been found to be H_2SO_4 -amine clusters (Cai et al., 2021). As expected,
400 the testing result showed dramatic underestimations for Beijing using model 1 with only H_2SO_4 concentrations
401 considered (Fig. 3(d1)), whereas models 2 (Fig. 3(d2)) and 3 (Fig. 3(d3)) yielded clearly enhanced J_5 predictions,
402 with relatively minor differences between models 2 and 3. These features are consistent with the fact that in addition
403 to H_2SO_4 , also other vapors are import to NPF and sub-5 nm particle growth in Beijing, and demonstrate that RH
404 and CS in our parameterization together determine in a complicated way the sources and sinks of these vapors, the
405 survival probability of sub-5 nm particles, and the potential emissions of sub-10 nm primary particles from traffic.
406 This study did not include amine-related compounds in the formulas because the lack of measured NH_3 data makes
407 the parallel comparisons difficult among the chosen sites for model training.

Commented [XL8]: Modification for RC1, Comment 3.

Also for RC2, Comment 4.

Commented [XL9]: Modification for RC1, Comment 4,5,6 combined. Part 1/6.

408 For Budapest, a large European city, the underestimates in modelled J_5 are not as much improved as they were for
 409 Beijing when including RH or CS in the parameterization, which is indicative of distinct particle formation
 410 pathways between Beijing and Budapest, even though both sites represent urban background environments. On
 411 one hand, it is worth noting that including RH (model 2, Eq. 3) resulted in a decrease in the correlation coefficients
 412 between the measured and modelled J_5 in Budapest from 0.54 to 0.46 (Table 3). This suggests that the role of RH
 413 in the NPF process in Budapest is less significant than other chosen inputs, despite previous indications that high
 414 RH levels have a strong potential to suppress NPF during non-event days in Budapest (Salma et al., 2021), even
 415 though the RH values in Budapest were considerably higher than those in Beijing (Table S2). On the other hand,
 416 including CS (model 3, Eq. 4) in addition to RH (model 2, Eq. 3) leads to an increase in the correlation coefficients
 417 between the measured and modelled J_5 from 0.46 to 0.61 (Table 3). We used both NPF and non-NPF days during
 418 model training even though it was found that CS was 50% lower during non-NPF events in Budapest than the
 419 values during NPF events (Salma et al., 2016). As a result, it is difficult to determine if the model's performance
 420 gain was entirely brought on by the addition of CS. Otherwise, the results are all in line with the earlier indirect
 421 evidence that chemical species other than H_2SO_4 influence the particle growth and possibly NPF process in
 422 Budapest (Salma and Németh, 2019). If one considers additional vapors other than H_2SO_4 for Budapest alone for J_5
 423 parameterization, one could include oxidation products of VOCs originating from either urban vegetation emissions
 424 or traffic emissions. For example, isoprene oxidation products can be used to describe the inhibiting effect on NPF
 425 (Heinritzi et al., 2020; Kiendler-Scharr et al., 2009), while monoterpene oxidation products could enhance sub-3
 426 nm particle growth (Kulmala et al., 2013).

Commented [XL10]: Addition text for RC2, Comment 7.

427 Based on the testing results, model 3 is more likely to predict a more accurate J_5 for Beijing based on the highest
 428 AIC ratio (except for 1), while model 1 predicts better for Budapest. Notice the fact that J_5 showed distinct levels
 429 of measured J_5 dependence with RH and CS in Beijing and Budapest (J_5 and RH: Beijing: $r = -0.21$, Budapest: r
 430 $= -0.1$; J_5 and CS: Beijing: $r = -0.02$, Budapest: $r = 0.57$, Fig. S1).

431 **4.1.3 Mediterranean rural site: Agia Marina, Cyprus**

432 For Cyprus, it appears that meteorology and condensation sink terms have only minor effects on the formation of
 433 5 nm particles in such rural environment under the influence of marine vapors when comparing the results from
 434 testing dataset across models (Fig. 3g). However, including RH improves slightly the correlation between the
 435 modelled and measured J_5 as seen in model 2 (r rises from 0.42 to 0.49, Table 3). The reduced values of r in model
 436 3 (Eq. 4) indicate a somewhat reversed impact of CS on J_5 , which requires additional examination as J_5 and CS are

437 weakly correlated ($r = 0.03$, Fig. S1). The H_2SO_4 concentration showed a low contribution to J_5 with the exponent
438 being less than 1 in model 4 (Table 3), which led to a more underestimated modelled J_5 comparing to model 1 (Fig.
439 3(g1), (g4)). This could be an indication that potentially other anthropogenic, biogenic or marine compounds are
440 of greater contribution to the particle formation processes in Cyprus than H_2SO_4 (Debevec et al., 2018). Owing to
441 the orographic conditions, the air mass types approaching to the Cyprus measurement site are mixed, including the
442 ones from North Africa, Marine, Europe, and northwest/southwest Asia. This results in the Mediterranean
443 atmosphere in Agia Marina containing various vapors that could influence NPF. The potential key contributors
444 could include oxidation products of dimethyl sulfide (DMS) originating from ocean plankton emissions (Rosati et
445 al., 2021), iodine oxidation products like HIO_3 (He et al., 2021), the stabilizing agent NH_3 (Jiang and Xia, 2017;
446 Lan et al., 2021; Lehtipalo et al., 2018; Yu et al., 2018), and oxidation products of VOCs from the surrounding
447 pinus forests and oaks under favorable meteorological conditions (Debevec et al., 2018).

Commented [XL11]: Addition text for RC2, Comment 8.

448 We should note that the measurements in Cyprus covered only two weeks in springtime, which limited our
449 quantitative observations in model training for other seasons compared to sites with long-term measurements.
450 Based on the findings above, model 1 seems to be the most suitable functional form for the prediction of J_5 in
451 Cyprus.

452 **4.1.4 Amazonian basin: Manacapuru, Brazil**

453 The measured and modelled values of J_5 from the Manacapuru site scatter around the 1:1 line in all the models
454 (Fig. 3). Previous studies reported high RH levels year-round in the measurement site near Manacapuru (Myers et
455 al., 2022; Zhao et al., 2022), which is expected to suppress NPF frequency and to lead to lower formation rates.
456 We observed such suppression effect when taking RH into account as shown by the increased correlation
457 coefficients from 0.004 to 0.19 (Table 3). Studies from Manacapuru suggested that the epoxide vapors could be a
458 potential precursor vapor in particle formation because of anthropogenic influences (Paulot et al., 2009), while Xu
459 et al., (2014) suggested the presence of epoxide vapors can enhance particle nucleation when RH levels increase.
460 We did not observe apparent improved model performance in model 3 when CS is included, as r remained almost
461 unchanged compared to model 2 (Table 3). One factor to consider is that we did not apply hygroscopic growth
462 factor when calculating CS for Manacapuru to maintain the consistency of the training dataset. However, the impact
463 of RH on CS, particularly on the actual particle surface area available for H_2SO_4 uptake, seems to be significant for
464 high RH environments like Manacapuru (Myers et al., 2022). Another assumption could be that even with the high
465 CS, it is still low enough to allow sufficient precursor vapors contributing to NPF processes.

These current findings provide evidence for H_2SO_4 being an effective enough precursor for the particle formation at 5 nm in the atmosphere of Manacapuru (model 1, Fig 3(h1)). However, the RH stabilization effect on H_2SO_4 is not exerted necessarily, as RH remains at high values at around $89 \pm 13\%$ despite whether it is measured during wet season or dry season (Myers et al., 2022). With these observations, model 1 with a focus on the H_2SO_4 concentrations manages to predict J_5 well for biogenic vapor dominated environment like Manacapuru.

Commented [XL12]: Modification for RC1, Comment 8.

5 Tracer model 5 simulation

We simulated the particle number size distribution (PNSD) in EC-Earth global chemical transport model TM5-MP (Tracer Model 5, Massively Parallel version, details in supplement) by applying it with our J_5 model 1 and 4. Together, we compared our simulation results with the acid-organic binary homogeneous nucleation model from Riccobono et al., (2014):

$$J_{\text{Riccobono}} = k_m \times [\text{H}_2\text{SO}_4]^p \times [\text{BioOxOrg}]^q, \quad (\text{Eq. 8})$$

where $k_m = 3.27 \times 10^{-21} \text{ cm}^6 \text{ s}^{-1}$, $p = 2$ and $q = 1$.

The details of the 14 tested measurement stations are shown in Table S7. Note that the data from these 14 stations are independent from any training or testing datasets used in the previous sections of this paper. Here, we essentially compared the simulated and measured PNSD in three particle modes (nucleation, Aitken, accumulation) from the entire year 2018 to assess the simulation accuracy among global environments.

Figure 4 shows the comparisons of PNSD between the on-site measurements and the TM5-MP simulations. For biogenic environments, simulations using model 4 shows the closest particle number size distribution to the measured ones particularly in Aitken mode particles, promoting the sulfuric acid-based nucleation mechanism involving the source-sink-meteorology even for environment dominated by biogenic vapors. For the Arctic region, model 4 simulated particle concentrations are overall overestimated, while model 1 simulation shows better alignment of particle number concentrations around the Aitken mode. This might indicate that the nucleation process has a lower dependence on the variations of meteorology than we expected. For coastal environments, even though Utö (Baltic Sea Island) and La Réunion (southern hemisphere island) are located at different hemispheres and also have different geographical settings, the nucleation mechanisms from models 1 and 4 both show similar predictions on particle concentrations across particle modes, with larger underestimation in the accumulation mode for model 1. This once again validates the source-sink-meteorology mechanism in model 4. By observing the ratio between the simulated and measured particle number concentrations, we can quickly see that the sulfuric acid-based particle formation mechanisms with (model 4) or without (model 1) meteorology inputs have successfully

narrowed the gap between the simulations and observations across all particle modes, with significant improvements for the nucleation mode (Fig. 5). The “Total” contains the simulated/measured particle number concentrations ratio from all particle modes, and it is obviously seen that applying model 4 improves the overall global PNSD simulation compared with the sulfuric acid-organic vapor binary model from (Riccobono et al., 2014). This observation shows that including the RH and CS is needed for better understanding of the global particle number size distributions.

6 Conclusion

The particle formation rate is one of the key characteristics in new particle formation studies. By utilizing distinct field measurement data, we can model the particle formation rate and estimate the overall atmospheric aerosol budget over different environments. We parameterized J_5 in four functional forms using the combined datasets from six environments, covering boreal forests (Hyytiälä, Värriö), urban sites (a megacity of Beijing and a large European city of Budapest) and rural environment (Cyprus, Manacapuru). The particle formation schemes involve the main precursor vapor H_2SO_4 , relative humidity (RH) and condensation sink (CS). Due to the small number of parameters and the diversity of environments included to generate the schemes, the roles of RH and CS are not only related to their potential direct impact on J_5 , but also to sources and sinks of vapors other than H_2SO_4 contributing to formation and growth of sub-5 nm particles and to potential emissions of sub-10 nm particles, e.g., from traffic. Overall, our models showed improved performances as RH and CS were taken into consideration. The model evaluations may suggest that particle formation mechanism is more sensitive to certain factors in specific environments. Sulfuric acid is an effective precursor vapor in NPF processes for most of the measurement sites we selected for model training. Nevertheless, relying solely on H_2SO_4 generally resulted in a weaker model performance for environments where the NPF schemes are dominated by biogenic emissions. This suggests that for developing globally applicable particle formation rate models, more precursor vapor types need to be included alongside H_2SO_4 . The purpose of the paper is twofold: first, to address the lack of knowledge regarding global particle formation rates for particles at 5 nm and larger, and second, to provide a globally applicable semi-empirical parameterization for the sulfuric acid-based neutral particle formation. The simplicity of the parameterization is demonstrated by three factors. First, NPF is a widespread occurrence in various types of environments, where the characteristics of particle formation share common mechanisms involving major precursor types and environmental factors. Second, the main input H_2SO_4 concentrations data can be obtained from field measurements or proxies, from which the

Commented [XL13]: Modification for RC1, Comment 4,5,6 combined. Part 3/6.

524 contribution of H₂SO₄ to NPF can be directly compared among global sites. Third, we skip the microphysics
525 complexity of sub-5 nm particles, where the physical and chemical properties differ significantly from those that
526 are above 5 nm when discussing particle formation and growth.

527 The limited data availability from certain sites (less than 1 year), such as Budapest, Cyprus and Manacapuru, should
528 be noted when applying our models. Conclusions drawn from these sites can be more confidently applied to the
529 specific seasons covered in the model training, such as springs being more representable for Budapest and Cyprus,
530 and summer to early autumn for Manacapuru.

531 Overall, our parameterization findings show that our models including H₂SO₄ concentration, RH and CS can predict
532 J_5 on a satisfactory level for various environment types at once. Among the tested models, models 3 and 4 (Eqs. 4
533 and 5) can be utilized for predicting J_5 in a global scale if (1) the H₂SO₄ concentrations are known whether through
534 field measurement or proxies, (2) the meteorology parameter RH is monitored continuously, and (3) the particle
535 number size distributions are sufficient and assessed to yield CS. Some caution should be maintained when utilizing
536 these models for environments with very low RH and/or high CS, especially if the high CS is related to primary
537 particle emissions, as the associations between these model parameters and J_5 are complicated and multifaceted.
538 While the parameterizations presented in this study offer an improvement over previous approaches, further
539 development is needed to incorporate vapors important for NPF, such as iodine oxoacids, particularly in marine
540 environments.

Commented [XL14]: Modification for RC1, Comment 4,5,6 combined. Part 6/6.

Commented [XL15]: Modification for RC1, Comment 4,5,6 combined. Part 4/6.

Commented [XL16]: Modification for RC1, Comment 9.

541 *Author contributions.* Measurements: NS, RB, CY, LQ, TJ, IS, MV, TW. Data Analysis: XL, LD, MZ, NS, RB,
542 CY, LQ, IS, PZ. Results interpretation: XL, LD, PP, TN. Discussions: all co-authors. Writing: XL, LD, PP, V-MK,
543 TN. Comments and revisions: all co-authors.

544 *Data availability.* The datasets used in this study are now available on Zenodo:
545 <https://zenodo.org/records/15295592>. The data for the 14 global measurement sites are from EBAS database
546 (<https://ebas-data.nilu.no/>, last access 13.12.2024).

Commented [XL17]: Datasets are now available on Zenodo.

547 *Conflict of interests.* At least one of the (co-)authors is a member of the editorial board of Aerosol Research.

548 *Code availability.* The MATLAB code used for the parameterization training in this paper is available on Zenodo:
549 <https://zenodo.org/records/15295592>.

Commented [XL18]: MATLAB code is now available on Zenodo.

551 *Acknowledgements.* [This work was supported by the European Research Council H2020 projects GASPARCON](#)
 552 [\(Grant No. 714621\), FORCeS \(Grant No. 821205\), and EMME-CARE \(Grant No. 856612\), by the Cyprus](#)
 553 [Government, by the Hungarian Research Development and Innovation Office \(contract: Advanced 150835\), by the](#)
 554 [Research Council of Finland with grant numbers 355330 and 360114, and by ACCC Flagship funded by the](#)
 555 [Research Council of Finland grant number 337549. L.D. received funding from the Swiss National Science](#)
 556 [Foundation \(Ambizione grant number 216181\). We are grateful to all the people who have contributed to the](#)
 557 [ambient measurements at measurement stations. The Finnish and Cypriot stations are part of the European research](#)
 558 [infrastructure ACTRIS. We acknowledge ACTRIS CiGas – Centre for Reactive Trace Gases In Situ Measurements](#)
 559 [and CAIS-ECAC – Centre for Aerosol In Situ Measurements for providing operational support to the aerosol and](#)
 560 [reactive trace gas instruments at the stations. We also thank Dr. James N. Smith for providing the GoAmazon2014/5](#)
 561 [data for Manacapuru.](#)
 562

Commented [XL19]: Added the missing grand number 821205.

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982 Figure 1. Map of measurement locations included in this study. The number markings indicate the exact locations
983 of the measurements. Created using a template from Canva (www.canva.com)

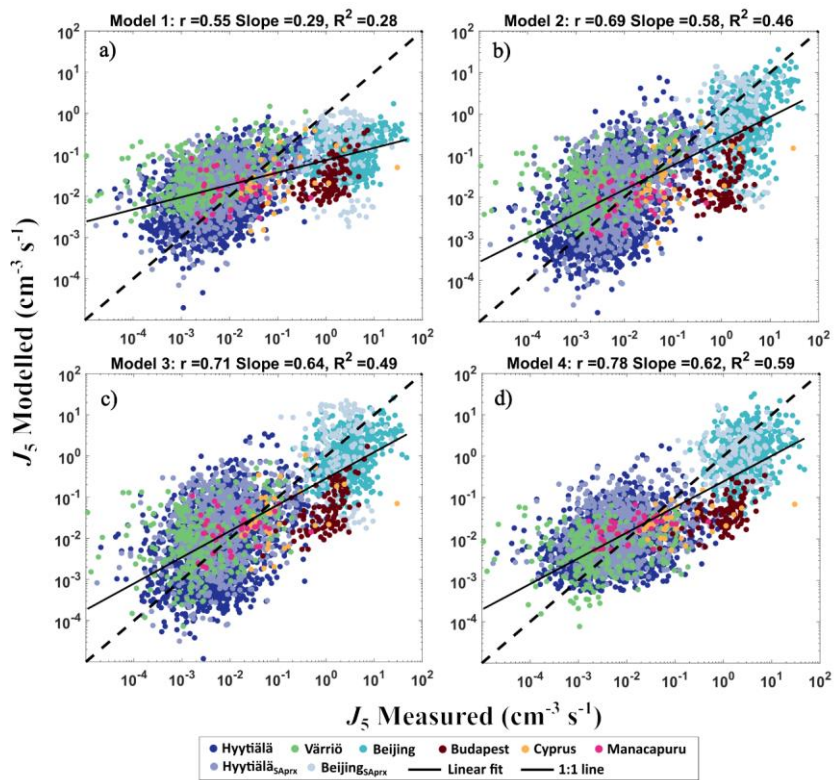
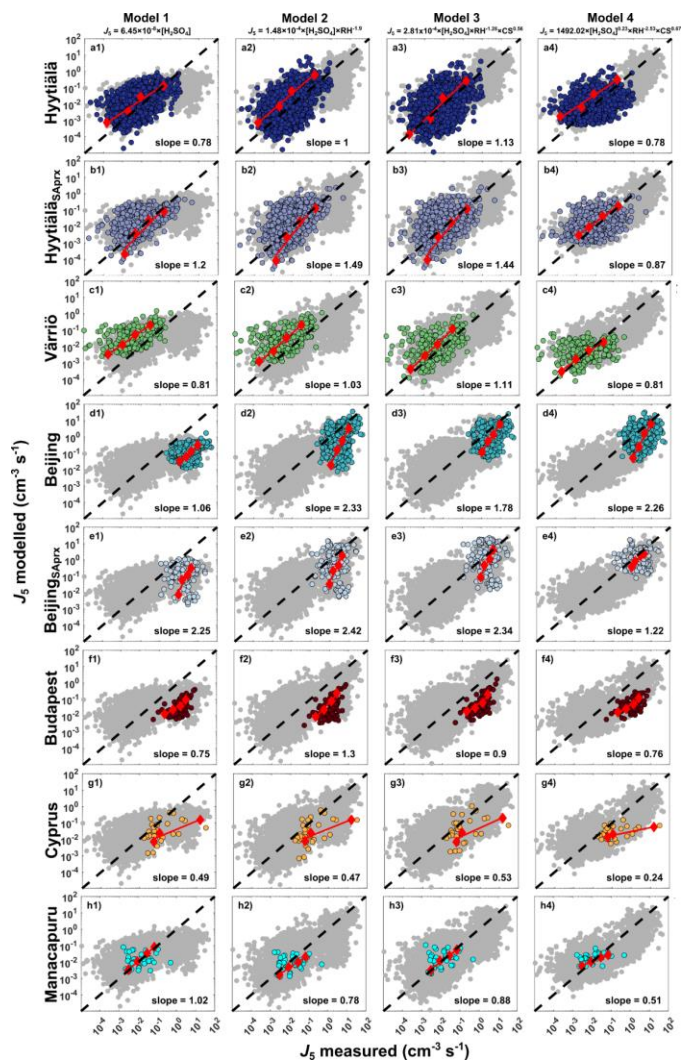


Figure 2. Modelled and measured J_5 scatterplots in logscale from four models using the testing dataset containing data from all sites in hourly time resolution. Each color represents the data from one measurement site, including datasets with H_2SO_4 proxy data from Hyytiälä and Beijing. The straight line showed the robust linear fit between the logscale modelled and the measure J_5 values, and the dashed line represented the 1:1 line. The correlation coefficient r , slope of linear fit, and the coefficient of determination R^2 are shown in the title of each subplot.



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993 Figure 3. Modelled and measured J_s scatterplots in logscale from four models using the testing dataset containing
 994 data from all sites in hourly time resolution. The labels on the left side of the y-axis are the site names. The

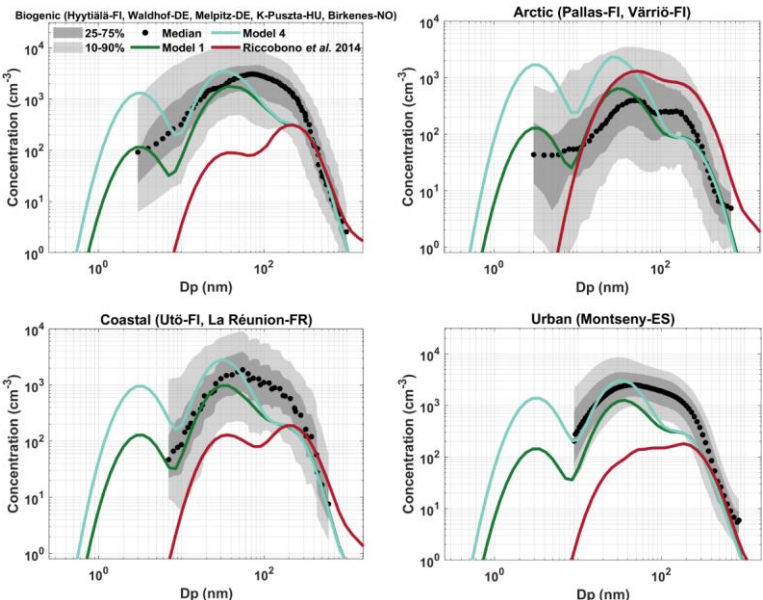


Figure 4. Environment-specific TM5-MP simulated particle number size distribution from 2018 in annual medians. Biogenic sites include rural and rural regional background environments (Hyytiälä, Waldhof, Melpitz, K-Pusztai and Birkenes); Coastal sites cover islands on the Baltic Sea and on the Indian Ocean in the southern hemisphere close to Madagascar (Utö and La Réunion); Arctic sites are two Finnish sites both situate within the Arctic Circle (Pallas and Värriö); Urban site is represented by a Spanish city Montseny.

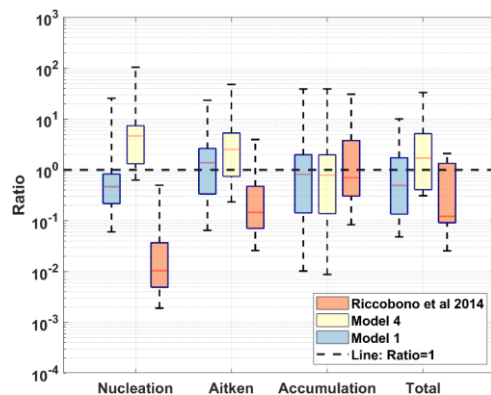


Figure 5. Ratio of simulated and measured particle number concentrations from the 14 global sites from TM5-MP simulations under three nucleation settings (Riccobono, model 1, and model 4) resulting in three particle modes (nucleation, Aitken, accumulation) in 2018 annual medians. The black line represents “ratio = 1” as a reference line. The “Total” represents the overall ratio between the simulation and the measurement particle number concentrations from all modes.

Table 1. Number of data points from each measurement site. The numbers in column “Total” account for the data points from six-site combined dataset utilized in model training and testing. The training set contains 75% of the total training data points, and 25% for testing set.

Sites	Hyytiälä	Beijing	Värriö	Budapest	Cyprus	Manacapuru	Hyytiälä _{proxy}	Beijing _{proxy}	Total
Training	5003	1342	728	367	140	140	-	-	7720
Testing	1642	501	248	109	34	40	797	164	3535

Table 2. Coefficient values (k_x , k_{RH} , k_{CS} , k_{SA}) retrieved from parameterization using training dataset. The term SSe represents the sum of squared error of each model. The units of k_x ($x = 1, 2, 3, 4$) vary as the functional form of model changes, while k_{RH} , k_{CS} , k_{SA} do not contain units. Since RH is counted using percentage (%), a dimensionless number, the scaling coefficients k_0 count mainly the units from H_2SO_4 concentrations and CS. As such, we must ensure that the RH input is in percentage.

Models	Functional forms	k_x	k_{RH}	k_{CS}	k_{SA}	SSe
1	$k_1 \times [H_2SO_4]$	$6.45E-8 \text{ (s}^{-1}\text{)}$				$2.78E+03$
2	$k_2 \times [H_2SO_4] \times RH^{k_{RH}}$	$1.48E-4 \text{ (s}^{-1}\text{)}$	-1.9			$5.16E+03$
3	$k_3 \times [H_2SO_4] \times RH^{k_{RH}} \times CS^{k_{CS}}$	$2.81E-4 \text{ ([s}^{-1}\text{]}^{0.45}\text{)}$	-1.28	0.56		$5.18E+03$
4	$k_4 \times [H_2SO_4]^{k_{SA}} \times RH^{k_{RH}} \times CS^{k_{CS}}$	$1492.02 \text{ ([cm}^{-3}\text{]}^{0.78} \times [\text{s}^{-1}\text{]}^{0.33}\text{)}$	-2.53	0.67	0.23	$3.36E+03$

Table 3. Summary of overall and site-specific correlation coefficients (r) four models using the testing dataset. The numbers in brackets under the site names represent the count for data points.

Slopes and r (robust linear fit), logscale										
	Models	Hyytiälä (1642)	Beijing (501)	Värriö (248)	Budapest (109)	Cyprus (34)	Manacapuru (40)	Hyytiälä _{SA} (797)	Beijing _{SA} (164)	Overall
Slope	1	0.43	0.33	0.32	0.58	0.35	0.04	0.34	0.07	0.30
	2	0.62	0.57	0.40	0.66	0.48	0.15	0.57	0.23	0.58
	3	0.48	0.43	0.32	0.85	0.37	0.18	0.42	0.12	0.64
	4	0.25	0.28	0.12	0.48	0.17	0.24	0.28	0.12	0.62
r	1	0.43	0.30	0.44	0.54	0.42	0.04	0.41	0.004	0.55
	2	0.47	0.32	0.47	0.46	0.49	0.19	0.48	0.07	0.69
	3	0.37	0.30	0.35	0.61	0.38	0.21	0.36	0.02	0.71
	4	0.31	0.22	0.18	0.51	0.37	0.46	0.33	0.09	0.78